

Methodology for the calculation of emissions from product usage by consumers, construction and services

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Methodology for the calculation of emissions from product usage by consumers, construction and services.

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Colophon

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Synopsis

Methodology for the calculation of emissions from product usage by consumers, construction and services.

This report describes the methods used to calculate emissions from product usage by consumers, construction and services by the Dutch PRTR for reporting greenhouse gas (GHG) emissions under UNFCCC, Kyoto Protocol, EU Monitoring Mechanism Regulation (MMR) and EU Effort Sharing Decision (ESD), and for international reporting obligations of other pollutants under CLRTAP and the NEC Directive.

Keywords: emissions, product use, air pollution, greenhouse gases

Publiekssamenvatting

Methodiekrapport om emissies door productgebruik te berekenen.

Het RIVM heeft beschreven hoe de Nederlandse Emissieregistratie de uitstoot berekent van verontreinigende stoffen door het gebruik van producten.

Nederland is vanwege internationale verdragen, zoals het Kyoto-protocol, de EU-Emissieplafonds (NEC-Directive) en de Convention on Long-range Transboundary Air Pollution (CLRTAP), verplicht om jaarlijks volgens de meest actuele wetenschappelijk inzichten te rapporteren over de uitstoot van broeikasgassen, verzurende stoffen en stoffen die gerelateerd zijn aan grootschalige luchtverontreiniging. De gerapporteerde uitstoot door productgebruik wordt met de methoden beschreven in dit rapport onderbouwd.

Deze rapportage is bedoeld voor de (internationale) reviewers die de Nederlandse rapportages aan de EU en VN valideren.

Kernwoorden: emissie, productgebruik, broeikasgassen, luchtverontreiniging

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1 Introduction

This document reports on the methods of calculating emissions caused by consumers and some small trade and service companies within the taskforce WESP (Taskforce on product usage by consumers, construction and services). Both greenhouse gas (GHG) and air pollutant emissions are calculated with these methods.

This document is only available online and is updated yearly if necessary. The emissions described in this document are part of the Dutch emission inventory (Dutch PRTR).

For more information, please check the website http://www.emissieregistratie.nl/ERPUBLIEK/bumper.en.aspx.

The emissions calculated by the WESP taskforce are largely caused by product uses which are mainly emitted to air. The emissions to the compartment water are calculated by the MEWAT taskforce. The emissions caused by industrial production, waste management and energy production are reported by the ENINA taskforce.

This document describes the background of the process causing the emissions as well as the method of data collection, the source of the emission factors and other important information. It is important to give a description of the method because, in some cases, the calculation is based on a single measurement or data from a single year. The base year needs to be corrected or interpreted to arrive at an estimation method for the total period.

One of the purposes of this document is to provide information on emission estimations. The WESP only calculates direct emissions caused by the process or product use; these are included in this report. Environmental effects such as acidification, greenhouse gas effect or ozone layer depletion are not considered. The waste produced, the amount of energy used, and the resources needed for production or the described process are not also included. Waste management and energy used are calculated on a country scale by the ENINA taskforce. The emissions caused by the production of the products used are calculated in other parts of the Dutch emission inventory, but only if production occurs in the Netherlands.

In the method description, we also explain how the spatial allocation occurs; this describes how the emissions are geographically distributed throughout the Netherlands. This is based on the location were the emissions are assumed to take place. For example, if a product is mainly used by consumers, the distribution is based on the number of people living in a certain area. The spatially distributed emissions are used as input for the (air quality) models calculating the concentrations of the substances in the environment. This is then used to get an estimation of the environmental quality in the Netherlands.

Appendix A gives an explanation of the quality indicators used in the tables in this report. The indicators are A till E, with A being the highest and E the lowest.

2 General assurance and quality control (QA/QC)

In accordance with the basic work agreements within the Dutch emission registration, the responsible work package leader checks that:

- 1. the basic data are well documented and adopted (check for typing errors, use of the correct units, and correct conversion factors):
- 2. the calculations have been implemented correctly;
- 3. assumptions are consistent and specific parameters (e.g. activity data) are used consistently;
- 4. complete and consistent data sets have been supplied.

Any actions that result from these checks are noted on an 'action list' by the ER secretary. The work package leaders carry out these actions and they communicate by e-mail regarding these QC checks, actions and results with the ER secretary. When adding a new emission year, the task forces perform a trend analysis in which data from the new year are compared with data from the previous year. The work package leader provides an explanation if the increase or decrease of emissions exceeds the minimum level of 5% at target group level, or 0.5% at national level. These explanations are also sent by e-mail to the ER secretary by the work package leaders The ER secretary keeps a logbook of all these QC checks and trend explanations and archives all relevant e-mails. This shows explicitly that the required checks and corrections have been carried out.

Based on the results of the trend analysis and the feedback on the control and correction process ('action list'), the Working Group on Emissions Monitoring (WEM) advises the institute representatives (Deltares on behalf of Rijkswaterstaat, Statistics Netherlands (CBS) and Netherlands Environmental Assessment Agency (PBL)) to approve the dataset. The ER project leader at RIVM formally approves the dataset on receipt of an e-mail by the institute representatives in which they give their approval. Furthermore, all changes of emissions in the whole time series as a result of recalculations are documented in CRF table 8(b).

3 Emissions of greenhouse gases

This report also provides the methodology descriptions of greenhouse gas emissions reported in the national greenhouse gas inventory. The relevant emission sources are presented in the following table, including a reference to the chapter and the CRF code.

CRF	Chapter	ES code	Emission source (English)	Emission source (Dutch)
1.A.4.b	21	0801801	Charcoal use for barbecuing	Houtskoolverbruik door consumenten: barbecuen
	25	T012200	Residential combustion, wood stoves and fire places	Vuurhaarden consumenten, sfeerverwarming woning
2.D.2	0	0801000	Burning candles	Branden van kaarsen
2.G.3.a	8	9310100	Solvent and other product use: anaesthesia	Oplosmiddel- en ander productgebruik: anesthesie, narcosegas
2.G.3.b	0	0811301	Solvent and other product use: sprays	Oplosmiddel- en ander productgebruik: spuitbussen, drijfgas/oplosmiddel, consumenten
2.G.4	18	0801700	Fireworks at New Year's Eve	Afsteken vuurwerk
	17	0850000	Degassing of groundwater, production of drinking water	Ontgassen drinkwater

Emissions from the use of compost by consumers are described in the methodology report for agriculture (Vonk et al, 2018: Methodology for estimating emissions from agriculture in the Netherlands)

In the next table, an overview is provided on the Tier used to calculate the emissions and the source of the emission factor.

GREENHOUSE GAS	CO ₂		CH₄		N ₂ O	
SOURCE AND SINK CATEGORIES	Method applied	Emission factor	Method applied		Method applied	
1. Energy						
A. Fuel combustion						
4. Other sectors	T1	D	T1	D	T1	D
2. Industrial processes						
D. Non-energy products from fuels and solvent use	T1	D				
G. Other product manufacture and use	T1	CS	T1	CS	T1	CS, D

D = default emission factors from the guidebook; CS = Country specific emission factor; T1 = Tier 1 methodology;

4 Emissions of air pollutants

This report provides the methodology descriptions of air pollutant emissions reported under the LRTAP convention in the Informative Inventory Report (IIR). The relevant emission sources are presented in the following table, including a reference to the chapter and the NFR code.

NFR	Chapter	ES code	Emission source (English)	Emission source (Dutch)
1A4bi	21	0801800	Barbecuing	Vleesbereiden: Bakken, braden en barbecuen
	21	0801801	Charcoal use for barbecuing	Houtskoolverbruik door consumenten: barbecuen
	25	T012200	Residential combustion, wood stoves and fire places	Vuurhaarden consumenten, sfeerverwarming woning
1B2aiv	24	8920900	NACE 47.3: gas stations, spills tank refill	SBI 47.3: Benzinestations, lekverliezen vullen autotank
	24	8920901	NACE 47.3: gas stations, vapour expel - tank refill	SBI 47.3: Benzinestations, verdrijvingsverliezen - autotanks
	24	8920902	NACE 47.3: gas stations, vapour expel - storage tanks	SBI 47.3: Benzinestations, verdrijvingsverliezen - opslagtanks
	27	8921100	NACE 46.71: wholesale trade in fuels and other mineral oil products	SBI 46.71: Groothandel in brandstoffen en overige minerale olieproducten
2D3a	15	0801100	Solvent and other product use: cosmetics	Oplosmiddel- en ander productgebruik: Cosmetica en artikelen voor persoonlijke verzorging. consumenten
	10	0802300	Solvent and other product use: car products	Oplosmiddel- en ander productgebruik: Autoproducten, consumenten
	31	0802400	Solvent and other product use: domestic pesticides	Oplosmiddel- en ander productgebruik: NMVOS huishoudelijke bestrijdingsmiddelen
	20	0802800	Solvent and other	Oplosmiddel- en ander

NFR	Chapter	ES code	Emission source (English)	Emission source (Dutch)	
			product use: leather maintenance products	productgebruik: Leer- en meubelonderhoud	
	7	0802901	Solvent and other product use: glues	Oplosmiddel- en ander productgebruik: Lijmen, consumenten	
	13	0803000	Solvent and other product use: detergents	Oplosmiddel- en ander productgebruik: Schoonmaakmiddelen, consumenten	
	20	0820600	Solvent and other product use: office products	Oplosmiddel- en ander productgebruik: kantoorartikelen, consumenten	
	36	0890401	Solvent and other product use: foam, applied in residential refrigerators	Oplosmiddel- en ander productgebruik: diffusie isolatieschuim koelkast/diepvriezer consumenten	
2D3d	0	0119800	Solvent and other product use: road-paint rural areas	Oplosmiddel- en ander productgebruik: Wegenverf buiten bebouwde kom	
	0	0129800	Solvent and other product use: road-paint urban areas	Oplosmiddel- en ander productgebruik: Wegenverf binnen bebouwde kom	
	0	0802200	Solvent and other product use: paint in construction	Oplosmiddel- en ander productgebruik: Verfgebruik bouw	
	0	0802201	Solvent and other product use: paint by consumers	Oplosmiddel- en ander productgebruik: Verfgebruik consumenten	
	6	0803100	Solvent and other product use: air fresheners	Oplosmiddel- en ander productgebruik: Luchtverfrissers, consumenten	
	0	8920800	NACE 45.2: specialised restoration of cars (painting and lacquering)	SBI 45.2: Gespecialiseerde reparatie van auto's (verven en lakken)	
2D3f	33	8922100	NACE 96.012: washing and (dry-) cleaning and dye- works (> 10 employees)	SBI 96.012: Chemische wasserijen en ververijen (> 10 werknemers)	
	32	8922200	NACE 96.012: washing and (dry-)cleaning and dye-	SBI 96.012: Chemische wasserijen en ververijen (< 10 werknemers)	

NFR	Chapter	ES code	Emission source (English)	Emission source (Dutch)
			works (< 10 employees)	
2D3i	23	0010300	Solvent and other product use: PCP pressure treated wood, stock	Oplosmiddel- en ander productgebruik: Emissie gevelbetimmering
	0	0801000	Burning candles	Branden van kaarsen
	28	0801001	Smoking cigars	Roken van sigaren
	28	0801002	Smoking cigarettes	Roken van sigaretten
	15	0801101	Solvent and other product use: cosmetics	Oplosmiddel- en ander productgebruik: Cosmetica en artikelen voor persoonlijke verzorging. HDO
	18	0801700	Fireworks at New Year	Afsteken vuurwerk
	10	0802301	Solvent and other product use: car products	Oplosmiddel- en ander productgebruik: Autoprodukten, HDO
	11	0802500	Solvent and other product use: carbolized wood	Oplosmiddel- en ander productgebruik: Gecarbolineumd hout, consumenten
	11	0802501	Solvent and other product use: carbolized wood	Oplosmiddel- en ander productgebruik: Gecarbolineumd hout, landbouw
	11	0802600	Solvent and other product use: carbol like wood preservatives	Oplosmiddel- en ander productgebruik: Gebruik carbolineum, consumenten
	11	0802601	Solvent and other product use: carbol like wood preservatives	Oplosmiddel- en ander productgebruik: Gebruik carbolineum, landbouw
	7	0802900	Solvent and other product use: glues	Oplosmiddel- en ander productgebruik: Lijmen, bouw
	13	0803001	Solvent and other product use: detergents	Oplosmiddel- en ander productgebruik: Schoonmaakmiddelen, HDO
	35	0804000	Solvent and other product use: creosote pressure treated wood, new	Oplosmiddel- en ander productgebruik: gecreosoteerd hout in de bouw, consumenten
	35	0804001	Solvent and other product use: creosote pressure treated wood, new	Oplosmiddel- en ander productgebruik: gecreosoteerd hout in de bouw, HDO

NFR	Chapter	ES code	Emission source	Emission source
			(English)	(Dutch)
	35	0804002	Solvent and other	Oplosmiddel- en ander
			product use:	productgebruik:
			creosote pressure	gecreosoteerd hout in
	25	0004002	treated wood, new	de bouw, landbouw
	35	0804003	Solvent and other product use:	Oplosmiddel- en ander
			creosote pressure	productgebruik: gecreosoteerd hout in
			treated wood, new	de bouw, verkeer en
			treated wood, new	vervoer
	35	0804100	Solvent and other	Oplosmiddel- en ander
			product use:	productgebruik:
			creosote pressure	opstand van
			treated wood, stock	gecreosoteerd hout in
		0001101		de bouw, consumenten
	35	0804101	Solvent and other	Oplosmiddel- en ander
			product use:	productgebruik:
			creosote pressure treated wood, stock	opstand van gecreosoteerd hout in
			liteated wood, Stock	de bouw, HDO
	35	0804102	Solvent and other	Oplosmiddel- en ander
		0004102	product use:	productgebruik:
			creosote pressure	opstand van
			treated wood, stock	gecreosoteerd hout in
			·	de bouw, landbouw
	35	0804103	Solvent and other	Oplosmiddel- en ander
			product use:	productgebruik:
			creosote pressure	opstand van
			treated wood, stock	gecreosoteerd hout in
				de bouw, verkeer en
	37	0011200	Industrial alconing	Vervoer
	37	0811200	Industrial cleaning of road tankers	Reinigen van tankauto's
	31	0812400	Solvent and other	Oplosmiddel- en ander
			product use:	productgebruik: NMVOS
			domestic pesticides	niet landbouw
		0000101		bestrijdingsmiddelen
	20	0820601	Solvent and other	Oplosmiddel- en ander
			products office	productgebruik:
	36	0890400	products Solvent and other	kantoorartikelen, HDO Oplosmiddel- en ander
	36	0070400	product use: foam	productgebruik: diffusie
			of refrigerators in	isolatieschuim
			waste dumps	koelkast/diepvriezer
				afvalfase
	29	8920700	NACE 45.1: service	SBI 45.1:
			stations, anti-	Garagebedrijven,
			corrosive treatment	antiroest beh.
	34	E800000	Solvent and other	Oplosmiddel- en ander
			product use:	productgebruik:
			fumigation of	ontsmetten transporten
			transports	

NFR	Chapter	ES code	Emission source (English)	Emission source (Dutch)
2H3	14	0802302	Building and	Stofemissies
			construction sites	bouwplaatsen
5C1bv	0	8922001	NACE 96.032:	SBI 96.032:
			crematories,	Crematoria, mortuaria
			mortuaries and cemeteries	en begraafplaatsen
5E	38	0801200	House fires	Woningbranden
	38	0801300	Car fires	Autobranden
	12	0890200	Solvent and other product use: scrapping of refrigerators	Oplosmiddel- en ander productgebruik: afdanken koelkast/diepvriezer
6A	19	0801600	Other sources and sinks: human transpiration and breathing	Transpiratie en ademen
	30	0802000	Manure from domestic animals	Huisdieren mest

5 Aerosol cans (CRF 2.G.3.b)

This section describes the emission of nitrous oxide from aerosol cans.

Process description	Emk_code	CRF_code	Sector
Aerosol cans	811301	2.G.3b	Consumers

5.1 Description emission source

Nitrous oxide (N_2O) is used as a propelling agent in aerosol cans (for example, cans of cream).

Contribution to the national emission

The contribution of this source to the total national N_2O emission was 0.82% in 2013.

5.2 Calculation

For the complete time series, emissions are calculated as follows:

Emission = Activity data x Emission factor

Activity data = Number of N_2O containing aerosol cans sold Emission factor = N_2O emission per aerosol can

a) Activity data

The Dutch Association of Aerosol Producers (NAV) reports data on the annual sales of N_2O -containing spray cans. Since the 2014 submission, the annual sales have been based on real sales figures instead of estimated sales. As a result of these improved activity data, the N_2O emissions have been recalculated for the whole time series.

b) Emission factor

The EF for N_2O from aerosol cans is estimated to be 7.6 g/can (based on data provided by one producer) and is assumed to be constant over time.

5.3 Uncertainty and Quality checks

For N_2O emissions, the uncertainty is estimated to be approximately 50 per cent based on expert judgement. Uncertainty in the activity data of N_2O use is estimated to be 50 per cent and that of the EF to be less than 1 per cent (the assumption is that all gas is released).

Quality checks

There are no sector specific quality checks performed. For the general QA/QC, see chapter 2.

5.4 Spatial allocation

The emissions of consumers are spatially allocated in the Netherlands based on population density.

Emission source/process	Allocation-parameter
Aerosol cans	population density.

Details available via

http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation)

5.5 References

The Dutch Association of Aerosol Producers (NAV)

5.6 Version, dates and sources

Version: 1.3

Date: September 2015

Responsibility: manager WESP task group

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6 Air Fresheners

This section describes NMVOC emissions from consumer air fresheners.

Process description	Emkcode	NFR code	Sector
Air freshener	0803100	2D3d	Consumers

6.1 Description of emission source

Air fresheners are used indoors to hider unpleasant odours and to fill the room with a pleasant scent. The ingredients include volatile organic compounds (VOCs) which enter the atmosphere as they travel from indoor to outdoor air. VOC Emission is calculated for the use of passive, electric and combustible air fresheners.

6.2 Calculation

Emissions from the use of air fresheners are calculated as follows:

Emission = Activity data x Emission Factor

Activity data = number of households in the Netherlands Emission factor = kg NMVOC emission from air fresheners averaged per household

a) Activity data

The number of households in the Netherlands are reported annually by the Dutch Central Bureau of Statistics (CBS) (www.cbs.nl).

b) Emission factor

The emission factor is calculated as sum of three separate emission factors derived for combustible, electric and passive air freshener products.

$$EF = EF_{combustible} + EF_{electric} + EF_{passive}$$

The emission factor per air freshener product is calculated as the emission rate of the product (g per hour) multiplied by the fraction of households using the products, the number of air fresheners present in the households that use the product, and the duration per year in which the product is used (hours per year).

 $\mathsf{EF}_{\mathsf{product}} = e_{\mathsf{NMVOC} \ \mathsf{per} \ \mathsf{product}} \ \mathsf{X} \ \mathsf{FR}_{\mathsf{households} \ \mathsf{using} \ \mathsf{AF} \ \mathsf{product}} \ \mathsf{X} \ \mathsf{N}_{\mathsf{AF} \ \mathsf{product} \ \mathsf{samples} \ \mathsf{per}}$ household $\mathsf{X} \ \mathsf{t}_{\mathsf{use}}$

The emission rate (g per hour) of an air freshener product is calculated by multiplying the weight of the product (g) by the weight fraction of the product that is NMVOC (g_{NMVOC} per g_{product}) divided by the product exhaustion time (hour), i.e. the duration at which all NMVOC has left the product.

 $e_{\text{NMVOC per product}} = (\text{Product Weight X WF}_{\text{NMVOC}}) / t_{\text{Exhaustion}}$

The VOC emissions per household are calculated per type of air freshener based on the use patterns described in the EPHECT ("Emissions, Exposure Patterns and Health Effects of Consumer Products in the EU") survey report (EPHECT, 2012), and product information of the most used brands of air fresheners reported in material safety data sheets (MSDSs). The EPHECT project is a European collaborative project co-funded by the European Union, in which important information has been gathered about the use of products by European consumers (EPHECT, 2015). The EPHECT survey was conducted in 2012 (EPHECT, 2012) and published in 2015 (Dimitroulopoulou et al., 2015a, b; Trantallidi et al., 2015); it includes survey data that describe the declared consumer use patterns of 4335 respondents including the use and non-use of air fresheners across Europe. Emission rates are calculated from the product information given in the MSDSs (SC Johnson, 2014; 2016a-e).

AF class	AF Products included	households using AF class using product ^A (%)	Product samples in households using products ^{A,B} (number)	Product Weight ^B (g)	Weight fraction NMVOCs (g per g) ^B	Product Exhaustion time (hour) ^B	Product use (hours per year) ^B	NMVOC Emission factor (g per year per household)
Combustible	Scented candles	22.6 ^A	1.7	96	0.025	27	219	4.4
Electric	Active evaporators	22	2.14	38.6	0.76	1440	4061	38.9
Passive	Passive evaporator	20	1.62	170	0.01	1440	8760 ^c	3.4
Total								46.7

A = see EPHECT, 2012

An emission factor of 46.7 g NMVOC per household per year has been calculated for the consumer use of air fresheners.

B = see Annex I

C = continuous emission (24 hours/day)

6.3 Uncertainty

Substance	Activity data	Emission factors	Emission
NMVOC	E	С	E

6.4 Spatial allocation

The emissions of consumers are allocated in the Netherlands based on population density. Details are available at

http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation).

6.5 References

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- SC Johnson. 2016a. Material Safety Data Sheet GLADE® CANDLE BLUE ODYSSEY. MSDS Number 350000022876
- SC Johnson. 2016b. Material Safety Data Sheet GLADE® CANDLE CASHMERE WOODS. MSDS Number 350000023246
- SC Johnson. 2016c. Material Safety Data Sheet GLADE® CANDLE HAWIIAN BREEZE. MSDS Number 350000027860
- SC Johnson. 2016d. Material Safety Data Sheet GLADE® CANDLE PURE VANILLA JOY. MSDS Number 350000023248
- SC Johnson. 2016e. Material Safety Data Sheet GLADE® 2 IN1 CANDLE SUNNY DAYS® & CLEAN LINEN. MSDS Number 350000023275
- Trantallidi, M., Dimitroulopoulou, C., Wolkoff, P., Kephalopoulos, S. and Carrer, P. 2015. EPHECT III: Health risk assessment of exposure to householdconsumer products. Science of the Total Environment. 2015, 536, 903-913

6.6 Version, dates and sources

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Date: January 2018

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6.7 Annex I: Emission rates derived from air freshener product information data

Table AI 1. Product information data: scented candles

Product Sample	Sample	Exhaustion	VOC Fraction
·	Weight (g)	Time (h)	(%)
Glade® 2 In 1 Candle -			
Sunny Days® & Clean			
Linen® - Jan 2014	96.3	30	1.5
Glade® 2 In 1 Candle -			
Vanilla Passionfruit &			
Hawaiian Breeze® - Aug			
2013	96.3	20	2.9
Glade® Candle - Apple			
Cinnamon - Jan 2014	96.3	30	0.6
Glade® Candle - Blooming			
Peony & Cherry™ - Aug			
2014	96.3	28	2.8
Glade® Candle - Blue			
Odyssey™ - June 2013	96	28	3.4
Glade® Candle - Hawiian			
Breeze® (Large Jar) - Feb			
2015	96.3	20	4.1
Glade® Candle - Pure			
Vanilla Joy™ - Jan 2014	96.3	30	4
Glade® Candle Cashmere			
Woods ® - Aug 2013	96.3	30	1
Average*1	96.3	27	2.54

^{*1} Average calculated from rows above

Table AI 2. Product information data: electric air fresheners

Product Sample	Sample Weight (g)	Exhaustio n Time (h)	VOC Fraction (%)
Glade Plugins® Scented Oil -			
Blooming Peony & Cherry™ - Oct			
2014	38.1	1440	84.4
Glade Plugins® Scented Oil -			
Clean Linen™ - Dec 2013	38.8	1680	72
Glade Plugins® Scented Oil -			
Hawaiian Breeze® - July 2014	37.9	1440	69.6
Glade Plugins® Scented Oil -			
Lavender & Vanilla - Dec 2013	37.4	1440	59.5
Glade Plugins® Scented Oil - Pure			
Vanilla Joy™ - Mar 2015	40.9	1440	95.2
Average*1	38.6	1440	76.14

^{*1} Average calculated from rows above

Table AI 3. Product information data: passive Air Fresheners

Product Sample	Sample	Exhaustion	VOC Fraction
	Weight (g)	Time (h)	(%)
Glade® Solid Air			
Freshener - Clean Linen™			
- Mar 2015	170	1440	1.5
Glade® Solid Air			
Freshener - Hawaiian			
Breeze® - Mar 2015	170	1440	0.5
Average*1	170	1440	1

^{*1} Average calculated from rows above

6.8 Annex II EPHECT Survey data

According to EPHECT, the average duration of scented candle being lit per occasion is, '85.6 minutes, ranging from 67 minutes in Spain and Italy to just over 120 minutes in Sweden.'

Table A II1. Average use frequency of combustible air fresheners calculated from EPHECT (2012) survey data

Multiple choice	% of	Frequency	(% of respondents)
answer	respondents	per year	X (Frequency per year)
At least once a day	19%	365	69.35
Several times a week	31%	104	32.24
Once a week	15%	52	7.8
Once every two weeks	9%	26	2.34
Once a month	9%	12	1.08
Less than once a month	12%	6	0.72
Weighted average (hours per year)			119.5

The average number of hours per day the consumer lights scented candles is then calculated as 85.6 minus X 153.4 per year = 219 hours per year.

Table A II2. Average use frequency of electric plug-in air fresheners calculated from EPHECT (2012) survey data

Multiple choice	% of	Frequency	(% of respondents)
answer	respondents	per year	X (Frequency per
			year)
At least once a day	40%	365	146
Several times a			
week	26%	208	54.08
Once a week	10%	52	5.2
Once every two			
weeks	3%	26	0.78
Once a month	7%	12	0.84
Less than once a			
month	8%	6	0.48
Weighted average (hours per year)			207

Table A II3. Average duration of electric air fresheners plugged in per occasion,

excluding the population who permanently leave the device on

Multiple choice	% of	Hours per	(% of respondents)
answer	respondents	plug-in	X (Hours per plug-in
		occasion	occasion)
Less than 1 hour	21%	0.5	0.105
Between 1 and 6			
hours	28%	3.5	0.98
Between 6 and 12			
hours	9%	9	0.81
Between 13 and 24			
hours	2%	19	0.38
Between 25 and 48			
hours	1%	36.5	0.365
Between 2 and 3			
days	1%	60	0.6
More than 3 days	1%	72	0.72
Weighted average (hours per year)		3.96	

The average use frequency of electric plug-in air fresheners is 207 times per year, whereas the average duration per occasion is 3.96 hours, excluding those consumers who permanently plug in the device. The number of hours per year is then 3.96 X 207 = 820 hours. However, the device is permanently plugged in at 37% of the households, so that on weighted average, the number of hours for all households in which an electric air freshener is plugged is 820 + 37% X 24 X 365 = 4061 hours

Table AII 4. The use of in-house air fresheners per room

Room	% of respondents claiming to
	have at least one electric
	freshener in the room
Living/dining room	60%
Bathroom	32%
Bedroom	30%
Hallway	29%
Kitchen	24%
WC	23%
Closet/Storage room	9%
Other room in the house	7%
Average number of in-house ^A air	
fresheners	2.14

A: calculated as the sum of the % of respondents claiming to have at least one electric freshener in the rooms, divided by 100%.

7 Adhesive products

This section describes the emissions of substances from the use of adhesives in consumer glue and the use of adhesives in the construction sector.

Process	EMK Code	Population	NFR category
description			
Adhesives	0802900	Construction	2D3i
		sector	
Adhesives	0802901	Consumers	2D3a

7.1 Description of the emission source

Volatile organic compounds (VOCs) are used as substance ingredients in adhesive products such as glues available on the consumer market, or adhesive products used in the construction sector. These VOCs are released to the air during the application of adhesive products.

7.2 Calculation

The total VOC emissions between 1990-2000 from adhesive products were estimated by interpolating monitoring data points (Table 1). The contribution of emissions from adhesives in consumer products and the construction sector is quantified according to the MilieuMonitor (1997) and KWS (2002) data. Annual emissions for the period after 2000 are set equal to the estimated emissions for the year 2000. Emissions of individual VOCs from adhesive use in consumer products and the construction sector are calculated as the total VOC emission from adhesives multiplied by the contribution of consumer products, or the contribution construction sector (Table 1) multiplied by the weight fraction of the individual VOC in the emission profile (Table 2).

Table 1. VOC emission for adhesives in consumer products and the construction sector

Year	VOC Emission	Contribution	Contribution	Data collection
ı oui	from	adhesives in	adhesives in	VOC emission
	adhesives	consumer	construction	
	(kt)	products (%)*A	sector(%)*A	
1981	3.95	19.6	36.0	Monitoring (CEA, 1994)
1990	3.84	19.6	36.0	Linear interpolation 1981-1991
1991	3.83	19.6	36.0	Monitoring (KWS, 2002)
1992	3.73	19.6	36.0	Linear interpolation 1991-1995
1993	3.62	19.6	36.0	Linear interpolation 1991-1995
1994	3.52	19.6	36.0	Linear interpolation 1991-1995
1995	3.41	19.6	36.0	Monitoring (KWS, 2002)
1996	3.37	18.7	34.0	Linear interpolation 1995-1998
1997	3.33	17.8	32.0	Linear interpolation 1995-1998
1998	3.30	17.0	30.0	Monitoring (KWS, 2002)
1999	3.30	17.0	30.0	Emission 1998
2000	1.58	17.0	30.0	Monitoring (KWS, 2002)

A: The remaining fraction of the total emission refers to emission from industrial sources

Table 2. Profile of individual substance weight fractions of VOC emissions from adhesives in consumer products and in the construction industry

Substance	ER-Nr	Weight fraction of VOC emission from consumer product	Weight fraction of VOC emission in construction
		adhesives	sector adhesives
propane	1031	0.099	0.070
isobutene	1042	0.099	0.070
Mixture C2-C10	1201	0.093	0.089
methyleenchloride	1303	0.036	
trichloro-ethane	1338	0.090	0.133
monohydroxy			
compounds	1629	0.047	0.051
esters kp <150 C	1653	0.093	0.103
propanon-2	1680	0.047	0.051
butanon-2	1681	0.233	0.256
dimethylether	1721	0.025	0.024
toluene	2502	0.140	0.153

7.3 Uncertainty and Quality

Substances	Activity data	Emission factors	Emission
VOC			D

7.4 Spatial allocation

Spatial allocation of emissions is based on population density.

7.5 References

KWS. 2002. KWS 2000 Eindrapportage Final Report
MilieuMonitor. 1997. Publicatiereeks emissieregistratie/MilieMonitor.
Nr. 37b, maart 2003: elektronische actualisatie: van Nr. 37 juli 1997.

7.6 Version, data and sources

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Date: January 2018

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8 Anaesthesia (CRF 2.G.3.a)

This section describes the emissions of N₂O used as an anaesthetic.

Process description	Emk_code	CRF_code	Sector
Anaesthesia	9310100	2.G.3.a	Trade and services

8.1 Description emission source

Nitrous oxide (N_2O) , commonly known as laughing gas, is still used as an anaesthetic. In addition to N_2O , other anaesthetics are in use in the Netherlands, most commonly halothane, desflurane, enflurane and sevoflurane. It is not known how much of these anaesthetics are used, and because of their decomposition, how much is exhaled again.

Contribution to the national emission

This emission is not a key source of greenhouse gases

The contribution of this source to the total national N_2O emission was <0.5% in 2009.

8.2 Calculation

For the complete time series, the emissions are calculated as follows: Emission = Activity data x Emission factor

Activity data = Amount of N_2O sold for anaesthesia Emission factor = N_2O emission from anaesthesia

This is a tier 1 methodology. The methodology is consistent with the IPCC 2006 Guidelines.

a) Activity data

The amount of nitrous oxide sold in the Netherlands would be the best measure for the activity data. Therefore since 2011, all companies known to sell nitrous oxide as anaesthesia to the Dutch market are asked to report their annual sales to the Dutch market. The total of these sales results in the total amount of nitrous oxide used as an anaesthetic in the Netherlands.

In the years prior to the reporting year (2010), an estimate was made based on the sales of the major supplier. In years where sales were not reported by all companies, the total was estimated based on the sales of the other companies; based on the estimated market share of the largest company selling nitrous oxide as an anaesthetic.

If nitrous oxide is sold as a mixture with oxygen, only the nitrous oxide is calculated as sales.

b) Emission factor

The emission factor for nitrous oxide sold in the Netherlands is 1kg per kg nitrous oxide. All nitrous oxide sold in a certain year is considered to be emitted after use in the same year.

This emission factor is consistent with the 2006 IPCC guidelines.

8.3 Uncertainty and Quality checks

In those cases where all companies report their sales, the uncertainty in the activity data is caused by stock changes on the consumer side (mainly hospitals). Those differences are considered to be negligible. If, on the other hand, not all companies provide their sales, the uncertainty can be as much as 25%. The uncertainty in emission factor is 0%, because all N_2O will be exhaled over time. Both the uncertainty in activity data and in emission factor are based on expert judgement.

Quality codes

Substance	Activity data	Emission factor	Emission
N ₂ O	В	Α	В

Quality checks

There are no sector specific quality checks performed. For the general QA/QC, see chapter 2.

8.4 Spatial allocation

Consumer emissions are spatially allocated in the Netherlands based on the number of beds per hospital.

Emission source/process	Allocation-parameter
Anaesthesia	Number of beds per hospital

Details available via

http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation)

8.5 References

Written (or in early years oral) data on sales from companies selling anaesthetics.

8.6 Version, dates and sources

Version: 1.2 Date: May 2015

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9 Burning Candles (CRF 2.D.2)

This section describes the emissions as a result of the burning of candles.

Process description	Emk_code	CRF_code	NFR_code	Sector
Burning of candles	0801000	2D2	2D3i	CON

9.1 Description emission source

Within households and in some catering industries, candles are burned to create a pleasant ambiance. Burning candles results in the emission of several substances, for example particulate matter and PAHs.

Contribution to the national emission

This emission is not a key source of greenhouse gases

The contribution of this source to the national emissions of benzo(ghi)pyrene is <10%. For benzo(a)pyrene, the contribution is about 1%.

9.2 Calculation

For the complete time series, the emissions are calculated as follows: Emission = Activity data x Emission factor

Activity data = Amount of candles burned in kg Emission factor = Emission per kg candle

This is a tier 1 methodology. The methodology is consistent with the IPCC 2006 Guidelines.

a) Activity data

The activity data consist of two parts: national statistics on the number of inhabitants of the Netherlands (Dutch bureau for statistics), and the amount of candles burned per person in the Netherlands.

The amount of candles burned per person is retrieved from a company selling candles in the Netherlands. Up to 2009 the amount of candles burned per inhabitant was retrieved from www.bolsius.nl. From 2010, it has been based on expert judgement by a representative from a company selling candles.

b) Emission factor

The emission factors for burning candles are dependent on the type of candle burned. Both tea lights and regular (gothic) candles have been taken into account, both estimated at 50% usage. Less-used candle types, e.g. beeswax candles, are not considered to be relevant for this calculation.

The CO_2 emission factor, given in g/MJ candle, is multiplied by a heating value of 42.7 MJ/kg. Both the CO_2 emission factor and the heating value for candles are derived from the Dutch fuel list 2015 (Zijlema, 2015). All other emission factors are mainly calculated based on EPA 2001.

Substance	EF	Unit
Benzo(ghi)pyrene	0.278	mg/kg candle
Benzo(a)pyrene	0.150	mg/kg candle
VOC	928	mg/kg candle
CO ₂	73.3	g/MJ candle
PM ₁₀	0.872	mg/kg candle
PM _{2.5}	0.872	mg/kg candle
Pb	1.56	mg/kg candle
Zn	0.127	mg/kg candle

9.3 Uncertainty and Quality checks

The uncertainty of both the activity data and the CO_2 emission factor are determined in the report on uncertainties in greenhouse gas emissions by Olivier (2009). The uncertainty in activity data is estimated to be 100%. The uncertainty in the emission factor for CO_2 is estimated at 20%.

For the other substances (not greenhouse gases), the uncertainty was not determined. Instead, the reliability of the data is qualitatively indicated in the table below with codes A-E (see Appendix A). The number of inhabitants in the Netherlands is accurately known, but the amount of candles burned per inhabitant is a rough estimate based on data from one manufacturer. Therefore, the activity data are relatively unsure and rated with a D.

The emission factors are retrieved by combining different sources to improve the reliability. However, since these sources did not take different candle types into account, the emission factors are rated with a C.

Quality codes

Substance	Activity data	Emission factor	Emission
Benzo(ghi)pyrene	D	С	D
Benzo(a)pyrene	D	С	D
VOC	D	С	D
CO ₂	D	С	D
PM ₁₀	D	С	D
PM _{2.5}	D	С	D
Pb	D	С	D
Zn	D	С	D

Quality checks

There are no sector specific quality checks; for the general QA/QC, see chapter 2.

9.4 Spatial allocation

The emissions of consumers are spatially allocated in the Netherlands based on the population density, following the assumption that most candles are burned in residential areas/households.

Emission source/process	Allocation-parameter
Burning of candles	Population density

Details available via

http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation)

9.5 References

- CBS, Dutch Bureau of Statistics, <u>Statline</u>, annual data on the number of inhabitants
- EPA; January 2001, Candles and incense as potential sources of indoor air pollution: Market analysis and Literature review; EPA-600/R-01-001.
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- Vebeka BV; Information on sizes and burning times:, http://vebeka.nl.
- Zijlema, P.J., 2015; Nederlandse lijst van energiedragers en standaard CO2 emissiefactoren.

9.6 Version, dates and sources

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10 Car Products

This section describes NMVOC emission from car products.

Process description	Emkcode	NFR code	Sector
Car products	0802300	2D3a	Consumers
Car products	0802301	2D3i	Trade and services

10.1 Description of emission source

NMVOC emission is estimated from the use of car products to maintain and clean company and private cars. Windscreen fluid is considered to contribute 70% of NMVOC emissions (ER, 2017), whereas other car products such as car wax, plastic cleaners, and cockpit sprays together comprise 30% of the emissions. Four different sources of emission are defined:

- 1. Windscreen fluid used on private cars.
- 2. Windscreen fluid used on company cars.
- 3. Other car products used on private cars.
- 4. Other car products used on company cars.

10.2 Calculation

Emissions from the use of windscreen fluids are calculated as follows:

Emission = Activity data x Emission Factor

Windscreen fluids

Activity data = total number of kilometres driven by automobiles in the Netherlands

Emission factor = kg NMVOC emission per driven kilometre in the Netherlands

a) Activity data

Data describing the number of driven kilometres by automobiles in the Netherlands are collected from the website of the Dutch Central Bureau for Statistics (CBS, 2017).

b) Emission factor

Data describing the market volumes of methanol, ethanol, and isopropanol as main ingredients in car windshield fluids in Finland for the years 2002-2014 (Table 1) became publically available in 2015 (ECHA 2015a,b,c). The market volumes of these substances in windscreen fluids sold in the Netherlands have been estimated from the Finnish market volumes by correcting for the number of kilometres driven in the Netherlands and Finland as well as the number of frost days in the Netherlands and Finland. Finland's climate is colder than the Netherlands, so relatively more windshield fluid is used for the purpose of de-icing. According to the USEPA, the use of windshield fluids at local temperatures above 0°C is about equal to 23% of windshield use at local temperatures below 0°C (USEPA, 1996; 2012). Finland has about

233 days of frost per year (Plantmaps.com), whereas the Netherlands only has 38 days of frost annually (KNMI, 2017). Therefore, the following correction factor for the difference in climate between Finland and the Netherlands has been applied:

$$\begin{split} correction_{climate} &= \frac{frost\; days_{Netherlands} + non\; frost\; days_{Netherlands} \times 0.23}{frost\; days_{Finland} + non\; frost\; days_{Finland} \times 0.23} \\ &= \frac{38 + 327 \times 0.23}{233 + 132 \times 0.23} = 0.43 \end{split}$$

Following on from this, the emissions of methanol, ethanol and isopropanol used in windscreen fluids per driven kilometre in the Netherlands are estimated as follows:

$$\frac{emission_{Netherlands}}{kilometers_{Netherlands}} = \frac{market\ volume_{Finland}}{kilometers_{Finland}} \times correction_{climate}$$

Table 1 includes the market volumes of methanol, ethanol and isopropanol in windscreen fluid in Finland (ECHA, 2015b), the number of kilometres driven by automobiles in Finland (Liikenneviraston Tilastoja, 2017) and the estimated emission per driven kilometre in the Netherlands (g/km).

	Kilometres driven in Finland (million km /y)	Finnish market volumes (t/y)		Estimated kilometre (g/km)		-	
Year		Methanol	Ethanol	Isopropanol	Methanol	Ethanol	Isopropanol
2002	31271	1326	3474	4323	0.02	0.05	0.06
2003	32211	1565	4061	4106	0.02	0.05	0.05
2004	33004	904	5606	3043	0.01	0.07	0.04
2005	33854	1334	4743	1995	0.02	0.06	0.03
2006	34473	1745	5061	2811	0.02	0.06	0.04
2007	34780	1358	5095	2617	0.02	0.06	0.03
2008	35661	1127	5952	1927	0.01	0.07	0.02
2009	35557	1246	6594	2892	0.02	0.08	0.03
2010	35868	1748	6353	1187	0.02	0.08	0.01
2011	36234	2559	7707	1746	0.03	0.09	0.02
2012	36740	935	4382	702	0.01	0.05	0.01
2013	36607	1819	6465	920	0.02	0.08	0.01
2014	36567	1422	4621	73	0.02	0.05	0.00
Average	34833	1468	5393	2180	0.018	0.066	0.028

The origin of the proportion of emissions from car products is assigned as follows: 60% from consumers and 40% by trades and services (ER, 2017). The emission factors for consumers and trades and services are calculated by multiplying the average estimated emission per kilometre in the Netherlands (g/km) over the period of 2002-2014 (Table 1) with these proportions (Table 2).

NMVOC in windscreen fluid	Emission factor consumers (g/km)	Emission factor trades and services(g/km)
Methanol	0.011	0.07
Ethanol	0.040	0.027
Isopropanol	0.017	0.011

Other car products

Activity data = amount of NMVOC in car products other than windscreen fluid sold in the Netherlands

Emission factor = emission per kg NMVOC in car products

a) Activity data

In 1997, bureau CREM conducted a study using car product monitoring data for the years 1994 and 1996 (CREM, 1997). The monitoring data originated from questionnaires filled in producers and suppliers of car products. In 1998, a recall survey was conducted by telephone in which the data for 1997 was established. The recall contained information from 26 companies (18 companies in the 1997 survey and 8 new companies). It was estimated that these 26 companies cover 80% of the market. The companies provided sales data of car products and the average amount of NMVOC these products contained. This information was then used to determine the NMVOC emissions from car products.

The 1997 monitoring data (CREM 1998) are still used for calculating NMVOC. The data from CREM (1998) is the most recent information available that refers to car products other than windscreen fluid.

b) Emission factor

NMVOC totals are recalculated to individual substances using an average car product profile. The profiles were established for car products by TNO (1992) in cooperation with the car products branch.

Substance in car product profile	factor
Propane	0.12
Isobutane	0.12
Monohydroxyverbindingen	0.54
Dimethyl ether	0.03
Hydrocarbon. mixture. c2-c10 <25% aromatic.	0.18

These totals include the emissions from windscreen fluids. Windscreen fluid is considered to contribute 70% of the NMVOC emission (ER, 2017). Therefore, the NMVOC totals have been multiplied by 30% to correct for this, so that only emissions from the use of car products other than windscreen are estimated. Furthermore, the emissions from other car products have been divided into the target groups of consumers ($\approx 60\%$) and of trade and services (mostly garages $\approx 40\%$).

10.3 Uncertainty

The uncertainty level of activity data for windscreen fluids is considered low, since the number of kilometres driven in the Netherlands is well known, therefore it is qualified with a B. The translation of the emission factors from Finland to the Dutch situation is more uncertain, therefore

qualified with a D. Both the emission factor and the activity data for the other car products are outdated, therefore the uncertainty is considered high and qualified with an E.

Since the NMVOC from windscreen products is about 70% of the emissions, the overall quality of the activity data is estimated with a D. The quality of the emission factors is qualified with a D.

10.4 Spatial allocation

The emissions of consumers and trade and services are allocated in the Netherlands based on population density. Details are available at http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation)

10.5 References

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10.6 Version, dates and sources

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Date: January 2018

Responsibility: manager WESP task group

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Note: Since 1999, no new data has become available on emission variables and emission factors of car products other than windscreen fluids. Therefore, the data used may be outdated.

11 Carbolineum treated wood

This section describes emissions resulting from Carbolineum treated wood.

Process description	EMK Code	Target group	NFR
Carbolineum treated wood	0802500	Consumers	2D3i
Carbolineum treated wood	0802501	Agriculture	2D3i
Carbolineum use	0802600	Consumers	2D3i
Carbolineum use	0802601	Agriculture	2D3i

Carbolineum is a pesticide that is brushed on wooden surfaces in order to prevent woodworm and moulds from affecting wooden constructions in housings, gardens and agriculture. The use of carbolineum for treatment of wood in contact with groundwater has been prohibited since 1999, and as a consumer product since 2001 (KWS, 2002).

11.1 Description of the emission source

Emission from carbolineum is released via two processes:

- Emissions from wood that has been treated with carbolineum in the past. Carbolineum contains VOCs and PAHs that evaporate to outdoor air or leach to soil from the surfaces of the treated wood.
- Emissions during the treatment of wood with carbolineum (VOC).

11.2 Calculation

The emissions of VOCs and PAHs from carbolineum treated wood are calculated by multiplying activity rates (AR) with respective emission factors (EF).

Emission = AR X EF

a) Activity data

Four different activity rates are considered:

- the area of carbolineum treated wood (treated in the past) for consumer use (m²)
- the area of carbolineum treated wood (treated in the past) for agricultural purposes (m²)
- the volume of carbolineum used by consumers during treatment (t.y⁻¹)
- the volume of carbolineum for agricultural purposes during treatment (t.y⁻¹).

It is assumed that 25% of the total volume of carbolineum used is for agricultural purposes and 75% for consumer use. It is also assumed that 25% of the total area of treated wood is for agricultural purposes and 75% for consumer. The total volumes of carbolineum used as well as the total surface area of carbolineum treated wood before 2001 are taken from Infomil (KWS, 2002). The use of carbolineum has been prohibited since 2001, so no use activity and no emissions since then

from treatment with carbolineum. The surface area of the carbolineum treated wood standing from earlier years is assumed to be reduced each year by 20%, which implies that 5 years after the prohibition of the carbolineum, there is no longer any carbolineum treated wood remaining, and therefore no longer any emissions occur from this source.

b) Emission factors

The emission factors for PAHs and VOCs released to the environment during treatment, as well as those for evaporation and leaching from the surface of treated wood is summarised in Table 1. The emission factors apply both to consumer use and agricultural use.

Table 1 Emission factors for carbolineum use and carbolineum treated wood.

	kg emission per million kg	kg emission per million m ² treated
Substance	carbinoleum used	wood
Naphthalene	0.02	0.0001
Anthracene	0.0005	0.00006
Fenanthene	0.015	0.0011
Fluoranthene	0.002	0.00049
Benzo(a)-anthracene	0.00027	0.000055
Chrysene	0.000026	0.00001
Benz(k)-Fluoranthene	2.9E-07	1.1E-07
Benzo(a)-pyrene	1.4E-06	5.70E-07
Benz(ghi)-Perylene	3.1E-07	1.20E-07
Pyrene	3.1E-07	1.2E-07
NMVOC	0.25	-

11.3 Uncertainty

Source	Activity data	Emission factors	Emission
Carbinoleum use	D	С	D
Carbinoleum treated wood	D	С	D

11.4 Spatial allocation

Spatial allocation of emissions is based on population density.

11.5 References

KWS 2000 eindrapportage, Infomil, Den Haag 52

11.6 Version, date, sources

Version: 1.0

Date: March, 2018 Task group leader:

Task group leader	ader Organization E-mail adress	
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12 CFCs from refrigerators and freezers

This section describes the emissions of chlorofluorocarbons (CFCs) as a result of the leakages in the refrigerant system and the processing of discarded refrigerators and freezers.

Process description	Emk_code	NFR_code	Sector
Discarding refrigerators	0890200	5E	Consumers
and freezers			

12.1 Description emission source

Since 1995, the production and sale of refrigerators and freezers (R/F) using chlorofluorocarbons as refrigerant has been prohibited in the European Union. However, given an average lifetime of at least 15 years, R/F equipment using CFCs is still in use, and significant numbers are discarded annually. In the Netherlands, discarded R/F equipment is collected and processed by specialised companies which remove and destroy the CFCs still present in the equipment. Still, in some cases the CFCs have leaked to the environment before the equipment is discarded and processed. This emission source represents these leakage emissions of CFCs and possible processing inefficiencies which lead to the emission of CFCs to the environment.

Before the year 2000, a share of the discarded R/F units were exported to Eastern Europe and Africa. However, from 1999, the export of old R/F equipment has been prohibited.

Contribution to the national total emission For trichlorofluormethane (CFC-12), the contribution of the discarded refrigerators and freezers is <25%.

12.2 Calculation

Emissions are calculated as follows: Emission = Activity data x Emission factor

Activity data = number of refrigerators and freezers that use CFCs as refrigerant that are discarded annually and that are not exported abroad. Emission factor = CFC emission per unit of CFC R/F equipment

a) Activity data

To estimate the number of CFC R/F equipment discarded annually from 1990 to 2030, a combination of multiple data is needed. First, the number of R/F units put on the market between 1960 and 1994 in the Netherlands was extracted from CBS data. These R/Fs are assumed to all use CFC-12 as refrigerant. The average amount of CFC refrigerant present per unit is estimated at 165 grams (Brouwer & Hulskotte, 1995).

The Weibull distribution is used to estimate the number of CFC R/F units discarded annually. Based on data from CBS, the average lifetime of a refrigerator is estimated at 16.4 years and of a freezer at 18.6 years.

Furthermore, the shape of the Weibull function is 2.2 for refrigerators and 1.3 for freezers (Magalini et al., 2014).

These calculations have resulted in an annual number of discarded CFC R/F units from 1990 to 2030. However, between 1990 and 1999, a share of the discarded units were exported and therefore did not cause emissions in the Netherlands. From 1990 to 1998, the share of discarded units was assumed to be 20%; in 1999 the share was assumed to be 10%. This ratio has been applied to the number of discarded CFC R/F units from 1990 to 1999 to estimate the number of CFC R/F units that were annually processed in the Netherlands. From 2000 onwards, the number of discarded CFC R/F units is assumed to be equal to the number of CFC R/F units processed.

b) Emission factor

The emission factor is estimated by subtracting the average amount of CFC refrigerant recovered from processed R/F units, from the average amount of CFC refrigerant used in CFC R/F units (Brouwer & Hulskotte, 1994).

165 gram - 60 gram = 105 gram/unit.

This calculation was verified by dividing the amount of recovered CFCs by the number of units processed according to the CFK 'aktieprogramma' (action programme) (1994) and subtracting this from the current amount of CFCs, which resulted in an emission factor of 103 grams/unit. Furthermore, the Flanders Environment Agency (MIRA, 2010) reports an average CFC recovery percentage of 33%, which results in an EF of 101 grams per unit.

Substance	EF	Unit
CFC-12	0.105	kg/RF unit

12.3 Uncertainty

The uncertainties of both the activity data and the emission factors have not been determined.

Only rough estimates are available for the number of CFC R/F units present, discarded or exported in the Netherlands. Therefore, the activity data has a significant level of uncertainty, rated D. The emission factor was verified by combining different sources, improving the reliability. It is therefore rated C.

Quality codes

Substance	Activity data	Emission factor	Emission
CFC-12	D	С	D

12.4 Spatial allocation

The emissions of consumers are spatially allocated in the Netherlands based on population density.

Emission	Allocation-parameter		
source/process			
Discarding refrigerators and freezers	Population density		

Details available from:

http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation)

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12.6 Version, dates and sources

Version: 1.0 Date: June 2015

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13 Cleaning Products

This section describes the emission from the use of cleaning products.

Table 1.

Process description	EMK Code	Target group	NFR category
Solvent and other	0803000	Consumers	2D3a
product use: detergents			
Solvent and other	0803001	Trades and	2D3i
product use: detergents		services	

13.1 Description of the emission source

VOCs are applied in cleaning products thanks to their fat dissolving capacities. The largest fraction of these VOCs is released to the air during or after the use of the cleaning products. The emissions described here are related to consumer use, institutional users and cleaning companies. Industrial cleaning products are considered by the ENINA task group.

Methylated spirit, spot removers, glass cleaner, hand cleaner, carpet cleaner and petrol are relevant product groups. A fraction of the methylated spirit (26%) is used as fuel, e.g. for barbecues and fondue sets. It is assumed that the VOCs used in fuels are completely combusted and thus not released. Medical hand disinfectants are included in the emission of VOCs from cleaning products. Ammonium is also classified as cleaning product. The ammonia dissolved in the ammonium can be released to the air during product use. Turpentine is another cleaning product for consideration, however emissions from this product are already included in the emission estimations for paint products.

13.2 Calculation

The VOC emissions from the use of cleaning products by consumers (Table 1) and trades and services (Table 2) were estimated by combining different data sources: (a) the monitoring data collected for the period 1990-1995 (Motivaction, 1996), (b) estimated emission volumes calculated by multiplying an emission factor by activity data for the period 1997-1999 (InfoMil, 2002) and 2000-2003 (NVZ, 2004; Nielsen, 2001).

Table 1. Monitored and calculated VOC emission from consumer use of cleaning products (Motivaction 1996; InfoMil 2004; NVZ, 2004)

	VOC emission per cleaning product (t/y)								Reference
Year	Methylated spirit	Spot remover	Window cleaner	Hand disinfectant	Air freshener	Carpet cleaner	Washing petrol	Total	
1990	1226	6	39	5	473	63	762	2574	Motivaction,
1991	1226	6	39	5	473	63	762	2574	1996
1992	1226	6	39	5	473	63	762	2574	
1993	1206	6	70	5	623	8	742	2660	
1994	1206	6	70	5	623	8	742	2660	
1995	1388	1	94	8	630	4	762	2887	
1996	-	-	-	-	-	-	762	2636	InfoMil, 2002
1997	-	-	-	-	-	-	762	2658	
1998	-	-	-	-	-	-	325	2211	
1999	-	-	-	-	-	-	325* ^A	2214	
2000	1177	200	185	58	662	16	325	2235	NVZ 2004;
2001	968	192	89	56	645	15	325	2290	Nielsen
2002	768	210	105	48	1699	12	325	3167	
2003	750	251	181	38	1654	27	325	3226	

^{*}A Emission from washing petrol is assumed to be equal to that of 1995

Table 2. Monitored and calculated VOC emission from the use of cleaning products in trades and services (Motivaction 1996; InfoMil 2004; NVZ, 2004)

	VOC emission per cleaning product (t/y)							Reference
Year	Methylated spirit	Window cleaner	Hand disinfectant	Air freshener	Carpet cleaner	Washing petrol	Total	
1990	0	156	1476	5	117	369	2123	Motivaction,
1991	0	156	1476	5	117	369	2123	1996
1992	0	156	1476	5	117	369	2123	
1993	0	344	1508	6	103	379	2340	
1994	0	344	1508	6	103	379	2340	
1995	0	178	1829	15	36	369	2427	
1996	-	-	-	-	-	3889	2636	InfoMil,
1997	-	-	-	-	-	3911	2658	2002
1998	-	-	-	-	-	3464	2211	
1999	-	-	-	-	-	3523	2270	
2000	198	1934	32	46	325	2536	198	NVZ 2004;
2001	184	1714	27	32	325	2283	184	Nielsen
2002	172	1684	173	16	325	2370	172	
2003	159	1701	212	19	325	2416	159	

^{*}A Emission from washing petrol is assumed to be equal to that of 1995

The Motivaction study also included the derivation of emission factors of water-dissolved VOCs in methylated spirit and in hand disinfectants. It was calculated that 38% of the VOCs in methylated spirit are released to the air, and that the fraction of VOCs in hand disinfectant released to the air is 50%. For the other products considered, it is assumed that 100% of VOCs are released to the air. The emissions after the year 1995 are calculated by multiplying these emission factors with the average amount of cleaning products used per household (NVZ, 2001) and the number of households (CBS).

Annual surveys were conducted by the NVZ (Nederlandse Vereniging van Zeepfabrikanten) amongst its members in the period 1996-2000. The emission of individual substance was determined by multiplying the total VOC emission volumes by an emission profile. This emission profile, designating the fractions of individual substances in the total emission of VOCs, was derived by TNO in cooperation with NVZ.

Washing petrol remains a complicated issue within this class of cleaning products because it is sold in several market streams, e.g. do-it-yourself product markets, so that the coverage of the Motivaction monitoring data from 1996 is low. Consequentially, the emission of VOCs from washing petrol has a high level of uncertainty. The annual emissions from washing petrol were estimated by InfoMil to be 1.5 kton for the period of 1990-1997 whereas the study performed in 1999 by AC Nielsen commissioned by InfoMil indicates 0.65 kton derived for the years 1998-2003.

Emission of the individual substances is calculated by multiplying the emission volumes given above by an emission profile expressing the weight fraction of the substance per ton emission of VOCs (Table 3)

Table 3. Profile of VOC emissions from cleaning products

VOC Substance	Weight fraction of VOC emission				
	from cleaning products (g/g)				
Propane	0.02284				
Isobutane	0.02284				
Methylene	0.005726				
Trichlor-ethane	0.004466				
Monohydroxy-components	0.700948				
Butanone2	0.009348				
Dimethylether	0.009559				

Ammonia is also considered to be a cleaning product. It is assumed that each household uses 1 litre of ammonia per year, so that the estimated emission is 1kton for the year 1990 (RIVM, 1994). The years following 1990 were corrected based on the number of households according to the CBS housing growth index.

The VOC emissions estimated for the years after 2003 are set equal to the emissions estimated for that year.

13.3 Uncertainty and Quality

Substance	Activity data	Emission factors	Emission
VOCs	D	D	D
Ammonia		D	D

13.4 Spatial allocation

Spatial allocation of emissions is based on population density.

13.5 References

CBS. http://statline.cbs.nl/statweb/publication/?vw=t&dm=slnl&pa=37 975&d1=0-5,10,15,19&d2=0&d3=0,5,10,15,(I-4)-l&hd=151214-1158&hdr=t&stb=g1,g2

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13.6 Version, data, sources

Version 1.0

Date: January 2018

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14 Construction sites

This section describes the emissions of particulate matter from construction sites.

Process	EMK code	NFR code	Sector	
description				
Building and	0802302	2H3	Trades and	
construction sites			services	

14.1 Description of emission sources

Particulate matter < 10 μ m (PM10) and < 2.5 μ m (PM 2.5) is released as fine dust as a consequence of different work activities at construction sites such as milling, drilling, grinding, chopping, sawing, and blistering rocks, bricks or concrete, as well as dust suspension from ground movement and heavy transport of debris and construction works.

14.2 Calculation

Emissions are calculated as follows:

Emission = Activity data X Emission Factor

Activity data = index number production revenues of the Dutch construction sector (1997=100)

Emission factor PM 10 = Reported emission of 10,620 kg per index point Emission factor PM2.5 = 3540 kg per index point

a) Activity data

The activity data refers to the production revenues of Dutch construction with the year 1997 as baseline with 100 index point. Index numbers for the years 1990-1999 are derived from CBS data that is unfortunately no longer available. The index numbers for the years 2000-2009 are derived from the prognostic reports of TNO (TNO, Bouw en Ondergrond 2005; 2007; 2008-2009) and the Dutch Ministry of Housing, Spatial Planning, and Environment (VROM, 2001; 2002), whereas the indexes for the years 2010-2017 are taken from the website of CBS (CBS, 2018). All index numbers are based on the production revenues of the Dutch construction sector expressed in currency (fl or €) corrected for economic deinflation and inflation.

b) Emission factor PM10

The emission factor for PM10 is calculated by dividing the estimated emission of 1062 t PM10 in the year 1997 (Haskoning, 2000) divided by the index number of 100. The 1062 t emission per year is calculated as the sum of the monitored PM10 emissions from (i) ground, water and road, (ii) civil and utility, (iii) demolition, and (iv) finish construction sites (Haskoning, 2000). As such, the emission factor for PM10 is 1062 t divided by 100 index point: 10,620 kg per index point.

c) Emission factor PM2.5

It is assumed that 1/3 of the weight of PM10 comprises PM2.5. As such, the emission factor of PM2.5 is equal to the emission factor of PM10 multiplied by 1/3: 3540 kg per index point.

14.3 Uncertainty

Substance	Activity data	Emission factors	Emission
PM10, PM2.5	С	D	D

14.4 Spatial allocation

Spatial allocation is based on population density.

14.5 References

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14.6 Versions, dates and sources

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15 Cosmetics for personal care

This section describes the emissions of NMVOC resulting from cosmetics and personal care products. The section does not describe the emission of NMVOC.

Process description	Emk_code	NFR_code	Target group
Cosmetics for personal care	0801100	2D3a	Consumers
Cosmetics for personal care	0801101	2D3i	Trade and services

15.1 Description emission source

Cosmetics products for personal care contain NMVOC which are emitted to the air during and after use. This includes products used by consumers and hairdressing salons, barbershops, and beauty parlours. The cosmetics for personal consist of a wide range of products: hairsprays, deodorants, eau de toilette/perfumes, nail polish/remover, aftershave and miscellaneous. In 2012, the emissions were ascribed to consumers for 96%, and to trades and services for 4%. This distribution shifted from 90% to 10% in 2004, to 96% and 4% in 2012. This is based on a communication from the NCV (Dutch Cosmetics Association; 2004 2012).

In 2004, hairspray was with ~65% the main NMVOC contributor of the cosmetics group, followed by deodorants ~29%. However in 2012, the two main contributors exchanged positions. Now deodorants contribute to ~54% (6.41 kton) and hairstyling products to ~35% (4.14 kton). According to the NCV, there have been no major changes of NMVOC concentration in products since 1996. Only within the group deodorant has there been an increase of the use of NMVOC-rich deodorant sprays. The increase of NMVOC emission is ascribed to the increased use of hairspray (flexible hairspray) and deodorant spray (roller sticks replaced by aerosol cans). Ultimately this has led to an increase in NMVOC from cosmetics.

Contribution to the national emission

The contribution of NMVOC by cosmetics for personal care is $\sim 33.6\%$ of the consumer NMVOC total, and 4.6% of the trades and services NMVOC total (ER 2011). The contribution of cosmetics to the Dutch total NMVOC emission is $\sim 7.7\%$ (ER 2011).

15.2 Calculation

The calculation of NMVOC emission cosmetics and personal care products is based on market share surveillance of these products, annually published by the NCV. In the past, the NCV themselves estimated the emission of NMVOC for the years 1997, 2002 and 2003 (NCV 1998, 2003 and 2004). After 2004, the market shares corrected with the annual Dutch central price index were used to estimate the emission.

In 2013, the NVC, on request of the ER, once again estimated NMVOC emissions from cosmetics and personal care products for consumers and trades and services (NVC 2013). Based on this latest NVC estimation, it can be concluded that the annual published NVC report contains sufficient information to estimate the NMVOC emission. The ER is estimated at 11.5 kton for consumers, while NVC were estimated at 11.7 kton. The 11.7 kton in 2012 will be used to index the emissions for the next 5 years. The members of the NCV represent the bulk of the market share in the Netherlands. The NMVOC total is, in turn, split up into the individual substances using an average profile established by TNO and the Dutch Cosmetics Association in 1992.

15.3 Uncertainty

The uncertainties of the emission calculation have not been quantified.

Quality codes

Substance	Activity data	Emission factors	Emission
NMVOC	В	С	С

15.4 Spatial allocation

The emissions of consumers and trade and services are allocated in the Netherlands based on population density. Details are available at http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation)

15.5 References

Annual reports NCV., www.NCV-cosmetica.nl

CBS statline: <u>www.statline.cbs.nl</u>

VOC emission cosmetics. Personal communication W.A. Pfeifer. Director

Dutch Cosmetics Association (NCV) 2004.

VOC emission cosmetics. Personal communication Dr. ir. R.T.H. van Welie technical director Dutch Cosmetics Association (NCV) 2013.

15.6 Version, dates and sources

Version: 1.1

Date: November 2013

Responsibility: manager WESP task group

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16 Crematoria

This section describes the emissions of dioxin, mercury and PM_{10} resulting from cremation of human remains. The section does not describe the emissions of NO_x , SO_x and CO_2 .

Process description	Emk_code	NFR_code	Target group
Crematoria	8922001	5C1bv	Trade and services

16.1 Description emission source

In the Netherlands there are 71 crematoria for human remains (LEV 2012). In 2011, 59% of the deceased were cremated.

Contribution to the national emission

Crematoria were responsible for the emission of 2.8% of the national mercury emission (ER 2012)

Crematoria were responsible for the emission of <<1% of the national dioxin emission (ER 2009)

Crematoria were responsible for the emission of <<1% of the national PM_{10} emission (ER 2009)

16.2 Calculation

Emissions are calculated as follows:

Emission = Activity data (-NeR) x Emission factor (-NeR) + Activity data (+ NeR) x Emission factor (+NeR)

Activity data (-NeR) = amount of cremations (in crematoria not yet in compliance with NeR)

Activity data (+NeR) = amount of cremations (in crematoria in compliance with NeR)

Emission factor (-NeR) = emission per cremation (in crematoria not yet in compliance with NeR)

Emission factor (+NeR) = emission per cremation (in crematoria in compliance with NeR)

a) Activity data

During cremation of human remains, substances are emitted; the most relevant substances are NO_x , SO_x , CO_2 , mercury, dioxin and fly ash (PM_{10}) .

Emission of NOx, SOx and CO2

The emissions are based on emission factors per cremated human. The emissions of the three substances were added to the other burning emissions of the target group trade and services. The ENINA task group is responsible for the calculation and reporting of the burning emission from trade and services.

Other emissions

The emissions are determined by multiplying the number of cremations in the Netherlands with emission factors. The number of cremations in the Netherlands is provided by the LVC (National Association Crematoria); data can be downloaded from the LVC home page.

Year	Deceased	Cremations		
	absolute	absolute	% total	% with NeR
1950	75,580	1,520	2	0
1980	114,279	39,947	35	0
1990	128,790	57,130	44	0
2000	140,527	68,700	49	5
2005	136,402	70,766	52	18
2010	135,895	77,465	57	75*
2011	135,516	78,599	58	86**

^{*} interpolation using year 2011

b) Emission factor

Mercury

The calculations of mercury are based on a study by Tauw Milieu (1997). Tauw made a list of amalgam sales in the past and combined it with the KUB model (1992). The KUB calculated an emission factor for mercury per age category. The emission factors are 1.15 in 1995, 1.37 in 2000 and 1.73 in 2010. In 2011, a factor of 1.73 g mercury per cremation was used. It is assumed that all the mercury in the amalgam is emitted to the air during cremation.

Implementation of the NeR measures

According to the cremations report ("crematies") (WESP 1996) 2150 m³ smoke-fume is generated per cremation resulting in 0.43 g Hg/cremation. However, measurements resulted in concentrations far below 0.43 g Hg/cremation (between 0.001 and 0.004 mg/m³). Assumption: emission of mercury with NeR = 0.05 mg/m3 \approx 0.1 g Hg/cremation.

Emission Hg (-NeR) (kg) = cremation (-NeR) x Hg emission factor (-NeR)

Emission Hg (+NeR) (kg) = cremation (+NeR) x Hg emission factor (+NeR)

Emission Hg total (kg) = Emission Hg (-NeR) (kg) + Emission Hg (+NeR) (kg)

PM10 (fly ash)

The emission factor for fly ash is **100g/cremation** (WESP 1996).

Implementation of the NeR measures

The NeR measures require the use of special filters (cloth or electrostatic filters) which reduce the emission of fly ash. According to WESP (1996), the emission for fly ash using cloth filters is 25 g/cremation. Measurements in Geleen showed concentrations of <6 mg/m³, or 13 g

^{**} calculation based on accurate list crematoria with NeR (LVC 2012)

per cremation. Measured data from Bilthoven showed even lower values, <0.7 mg/m³.

Assumption: emission fly ash with NeR = 10 g/cremation.

Emission fly ash (-NeR) (kg) = cremation (-NeR) x fly ash emission factor (-NeR)

Emission fly ash (+NeR) (kg) = cremation (+NeR) x fly ash emission factor <math>(+NeR)

Emission fly ash total (kg) = Emission fly ash (-NeR) (kg) + Emission fly ash (+NeR) (kg)

Dioxins

The emission factor for dioxins is 4 ug I-TEQ per cremation. The 4 ug I-TEQ/cremation is based on measurements made in 1991 of the smoke fumes from three crematoria (Bremmer et al, 1993).

Implementation of the NeR measures

For dioxins, the emissions with NeR measures are also lower. Measurements performed by TNO showed 0.024 $ng/m^3 = 0.052$ ug I-TEQ/cremation at Geleen and 0.013 $ng/m^3 = 0.028$ ug I-TEQ/cremation at Bilthoven.

According to measurements by the EEFS (European branchorganisation) lower values are possible. 0.1 ng/m³ (or 0.2 ug I-TEQ/cremation) is the modern German limit (27e BlmSchV) for installation with filters.

Assumption: emission dioxins with NeR = 0.2 ug I-TEQ/cremation.

Emission Dioxins (-NeR) (kg) = cremation (-NeR) x Dioxin emission factor (-NeR)

Emission Dioxins (+NeR) (kg) = cremation (+NeR) x Dioxin emission factor (+NeR)

E. Dioxins total (kg) = Dioxin emission (-NeR) (kg) + Dioxin emission +NeR) (kg)

Measures influencing the calculation

Since July 1998, new crematoria or new ovens in existing crematoria must comply to the NeR with respect to mercury (0.2 mg/m³). Therefore, the emissions of dioxins and fly ash are lower. According to the LVC information (2012), 57 of the 71 crematoria now comply with the NeR. The 57 crematoria with NeR accounted for 86% of the cremations. At the end of 2012, all the crematoria must be incompliance with the NeR. From this date onwards, only the lower emission factors are used.

16.3 Quality codes

Substance	Activity data	Emission factors	Emission
Dioxins	Α	В	В
Fly as	А	В	В
Mercury	Α	В	В

16.4 Spatial allocation

The emissions of the crematoria are assigned to the locations of the crematoria (SBI 96.032) in the Netherlands according to the ratio of employees.

Details available via

http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation)

16.5 References

ER, 2009. Netherlands Emission Registry. Data on 2009 available from www.emissieregistratie.nl

NeR op www.infomil.nl, november 2006

LEV 2012. http://www.lvc-online.nl/

RIVM, 1993, Bremmer, H.J, et al, november 1993, Emissies van dioxine in Nederland, RIVM, rapportnr 770501003, Bilthoven

Tauw Milieu 1997, Onderzoek naar kwikemissies van crematoria en beschikbare RGRtechnieken, B Hoekstra, R3517616 Tauw Milieu, Deventer, mei 1997

TNO 1996, Smit, E.R., 1996, Massabalans en emissies van in Nederland toegepaste crematieprocessen, TNO-rapport TNO-MEP-R96/059.

WESP 1996, J.G. Elzenga, WESP document Crematoria (H7), 1996.

16.6 Version, dates and sources

Version: 1.1

Date: November 2013

Responsibility: manager WESP task group

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17 Degassing of groundwater (CRF 2.G.4)

This section describes emissions from the degassing of groundwater.

Process description	Emk_co	CRF_cod	Sector
	de	е	
Degassing groundwater	0850000	2.G.4	Drinkwater
			companies

17.1 Description emission source

A part of the Dutch drink water is obtained through the production of ground water. Gasses dissolved in the water are released during processing, including methane. Shallow groundwater extraction for use in agriculture or on construction sites is not included in this document, as this water contains no methane.

Contribution to the national emission

This emission is not a key source of greenhouse gases

The contribution of this process to the national methane emission was <0.3% in 2009.

17.2 Calculation

For the complete time series, the emissions are calculated as follows: Emission = Activity data x Emission factor

Activity data = Amount of groundwater produced Emission factor = Emission per m³ groundwater

This is a tier 1 methodology. The methodology is not provided within the IPCC 2006 Guidelines.

a) Activity data

The amount of groundwater extracted for drink water purposes is used as activity data. The data are retrieved from VeWin, the Dutch society for (drink) water producing companies. Only the amount of groundwater is taken into account.

In the year 2009, this resulted in 676 million m³.

b) Emission factor

The emission factor for degassing groundwater was calculated for the year 1990 by dividing the estimated methane emissions (2000 tons) by the amount of extracted groundwater (810 million m³), as reported by van den Born (1990).

Substance	EF	Unit
CH₄	2469	kg/ million m ³ groundwater

17.3 Uncertainty and Quality checks

The activity data for the degassing of drink water from ground water are derived from the statistics of VeWin (Dutch association for water win companies). It is estimated that the uncertainty is 10% at most, based on expert judgement.

The uncertainty of the emission factor for methane (50%) is derived from 'Olivier, 2009'.

Quality codes

Substance	Activity data	Emission factor	Emission
CH ₄	Α	D	С

Quality checks

No sector specific quality checks were performed. For the general QA/QC, see chapter 2.

17.4 Spatial allocation

Consumer emissions are spatially allocated in the Netherlands based on population density. Although this might not be completely accurate, it may be the best assumption.

Emission source/process	Allocation-parameter
Degassing of groundwater	Population density

Details available via

http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation)

17.5 References

Born, G.J. van den, A.F. Bouwman, J.G.J. Olivier, R.J. Swart, 1991, The emission of greenhouse gases in the Netherlands, RIVM-rapport 222901003.

VeWin, vereniging voor waterwinbedrijven. http://www.vewin.nl/

17.6 Version, dates and sources

Version: 1.2 Date: May 2015

Responsibility: manager WESP task group

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18 Fireworks (CRF 2.G.4)

This section describes the emissions from fireworks. This refers specifically to consumer fireworks set off on New Year's Eve, since only professional fireworks are permitted on other occasions. Other amounts are considered to be negligible.

Process description	Emk_code	CRF_code	NFR_code	Sector
Fireworks	0801700	2.G.4	2D3i	CON

18.1 Description emission source

On New Year's Eve, inhabitants of the Netherlands are allowed to set off fireworks. These include both 'firecrackers' and 'ornamental' types of firework, with an estimated 15% / 85% ratio. When setting off fireworks, various gasses and metal substances are emitted, depending on the type of firework.

Contribution to the national emission
This emission is not a key source of greenhouse gases

The contribution of fireworks to the total national emission of particulate matter was <5% in 2009. However, the emission of some metal components contributes considerably to the national total. Fireworks may contain, among other substances, antimony, barium and strontium. For these metals, fireworks are the main source of process emissions, emissions to the soil, and emissions to the sewer system.

18.2 Calculation

The emissions are calculated as follows: Emission = Activity data x Emission factor

Activity data = Amount of fireworks lighted in kg Emission factor = Average emission per kg fireworks

This is a tier 1 methodology. The methodology is not provided within the IPCC 2006 Guidelines.

a) Activity data

In 2009, an estimated 17.3 million kg of fireworks were set off on New Year's Eve.

In order to calculate the amount of fireworks, the difference between import and export each year is taken into account. These statistics are derived from the national statistics agency (CBS). As the difference between imported and exported fireworks does not need to be set off in the same year, consecutive years have been averaged according to the following rule:

Year2 = (Year1 + 2 * Year2 + Year3) / 4

There is a bias in the statistics, since smaller companies are not included and the import of illegal fireworks is not accounted for. To compensate for this bias, the amount calculated from the statistics is multiplied by a factor 1.7. This is based on expert judgement (estimated total fireworks divided by the CBS reported amount). Prior to 1996, the CBS also provided information on the smaller companies, therefore in those years; the statistics were multiplied by 1.316 to correct for illegal fireworks.

b) Emission factor

The emission factors for all gaseous substances caused by fireworks are derived from Brouwer et al. (1995). This includes; CO_2 , CO, CH_4 , H_2S , SO_2 , N_2O and PM_{10} .

Substance	EF	Unit	
CO ₂	43250	kg/ million kg fireworks	Brouwer et al, 1995
CO	6900	kg/ million kg fireworks	Brouwer et al, 1995
CH ₄	825	kg/ million kg fireworks	Brouwer et al, 1995
H ₂ S	1195	kg/ million kg fireworks	Brouwer et al, 1995
SO ₂	1935	kg/ million kg fireworks	Brouwer et al, 1995
N ₂ O	1935	kg/ million kg fireworks	Brouwer et al, 1995
PM ₁₀	14244	kg/ million kg fireworks	Brouwer et al, 1995

For the emissions of heavy metals, studies by Brouwer et al. (1995), Plinke et al. (2001), both for Germany and Sweden, and Croteau et al. (2010) were combined to derive an average emission factor. The metal content described by Plink et al (2001) is a metal content per kg pyrotechnic mixture. To calculate a metal content per kg fireworks, it is assumed that fireworks consist of 40% pyrotechnic ingredients (comparable to Brouwer et al.).

The following table shows the metal emission factors of 'ornamental fireworks' from the several studies and the emission factor that is currently used for emissions from 'ornamental fireworks' in the Netherlands. The values in this table are based on the metal content of the fireworks (in g/kg fireworks) and form an emission factor for pollutants to all compartments (atmosphere, surface water, sewer and soil).

Pollutant	Brouwer 1995	Plinke 2001 (Germany)	Plinke 2001 (Sweden)	Croteau 2010	EF used in NL
Strontium	7.74	2.48	1.16	5.96	4.33
Barium	22.61	6.71	15.60	8,85	13.44
Copper	7.43	0.34	3.08	16,30	6.79
Antimony	0.92	0.01		2,34	1.09
Zinc			0.52	0.84	0.68

There is a difference between firecrackers and ornamental fireworks; the uncoloured 'fire crackers' do not contain these heavy metals and have a different emission profile. Therefore, the emission factors are weighted for the contribution of the different types of fireworks, with a ratio of 15% firecrackers and 85% ornamental fireworks.

The emission of particulate matter is calculated by the sum of the emitted metals and the other particulate matter forming components, as reported by Brouwer et al. (1995). According to Brouwer (1995), only 10% of particulate matter (and likewise the heavy metals) is emitted to

air. This is in-line with Croteau et al. (2010), where both the composition of the fireworks and the emissions to air are reported, giving around 10% of the ingredients as emissions to air. However, if we exclude the non-combusted residue from fireworks, then the fraction to air is much higher, between 20% and 48%. Other information sources suggest that the atmospheric emissions could be 3 times higher, which means that 30% of the emissions would be emitted to the air. The atmospheric emissions will be updated next year. This year, we have used a fraction of 10% of the emissions.

The remaining 70% of the emissions are emitted to sewer and soil. It is assumed that 14% is emitted to the sewer system and 56% is emitted to the soil.

The amount of metals that enter the sewer system is estimated based on the amount of paved area and the amount of precipitation entering the sewer system. In 2012, the amount of paved and sewer area was 4.29 * 10⁹ m². This includes both built-up areas and traffic areas in the Netherlands. As fireworks are mainly used in built-up areas and on the streets in cities on New Year's eve, it is assumed that the emissions from fireworks are released in the built-up areas and the traffic areas. With an average amount of precipitation of 850 mm per year, this would result in 3.65 * 109 m³ precipitation falling on the paved area. In 2015, the sewer system received an influent of 0.58 * 10⁹ m³ of precipitation (estimation by Liefting et al, 2017). This implies that only a fraction (0.58 / 3.65 = 16%) of the precipitation on paved area enters the sewer system. In the study by Liefting (2017), approximately 20% of the influent to waste water treatment plants is not accounted for by waste water from companies and households and by precipitation. This fraction could enter the sewer system via other routes, but it is also possible that the estimation of the amount of precipitation in the sewer system is too low. If we assume that 10% of the water that is not accounted for is also precipitation, then 0.58 * $10^9 + 0.20 * 10^9 = 0.78 * 10^9 \text{ m}^3$ precipitation enters the sewer system. This would result in a fraction of 0.78 / 3.65 = 21% of the precipitation on paved areas that enters the sewer system.

On average, we assume that 20% of the precipitation on paved areas enters the sewer system. Most of the fireworks are used in built-up areas and on traffic areas in cities, which means that the main part of the emissions also falls on the paved areas. We assume that the fraction of fireworks that enter the sewer system is equal to the amount of precipitation that enters the sewer system. Thus, 20% of the emissions on the ground is assumed to enter the sewer system. The remaining 80% of the emissions that falls to the ground is assumed to emit to soil.

This result in an emission fraction of 70% * 20% = 14% to the sewer system and 70% * 80% = 56% to soil.

This results in the following emission factors to the several compartments (kg/million kg fireworks):

Substance	Atmosphere	Sewer system	Soil
CO ₂	43250		
CO	6900		
CH ₄	825		
H ₂ S	1195		
SO ₂	1935		
N ₂ O	1935		
PM ₁₀	14244		
PM _{2.5}	14244		
Strontium*	433	606	2425
Barium*	1344	1882	7526
Copper*	679	951	3802
Antimony*	109	153	610
Zinc*	68	95	381

18.3 Uncertainty and Quality checks

In the Netherlands, the emissions of fireworks and candles are reported on a aggregated level under CRF 2G. For this aggregated level, Olivier (2009) reported uncertainties for CO_2 (20%), CH_4 (50%) and N_2O (50%). The uncertainty in activity data for fireworks is estimated to be 50% (Olivier, 2009).

The uncertainty of the emissions of other substances have not been studied. Instead, the reliability of the data is qualitatively indicated in the table below with codes A-E (see Appendix A). The codes are based on expert judgement.

Quality codes

Substance	Activity data	Emission factor	Emission
CO ₂	D	D	D
CO	D	D	D
CH ₄	D	D	D
H ₂ S	D	D	D
SO ₂	D	D	D
N ₂ O	D	D	D
Strontium	D	С	D
Barium	D	С	D
Copper	D	С	D
Antimony	D	С	D
Zinc	D	С	D
PM ₁₀	D	D	D
PM _{2.5}	D	D	D

Quality checks

No sector-specific quality checks were performed. For the general QA/QC, see chapter 2.

18.4 Spatial allocation

The emissions of consumers are spatially allocated in the Netherlands based on population density.

Emission source/process	Allocation-parameter
Fireworks	Population density

Details available via

http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation)

18.5 References

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18.6 Version, dates and sources

Version: 1.2 Date: May 2015

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19 Human ammonia emissions from transpiration and breathing

This section describes the emissions of NH₃ from human transpiration and breathing.

Process description	Emk_code	NFR_code	Sector
Human transpiration	0801600	6A	Consumers
and breathing			

19.1 Description emission source

This emission source describes ammonia emissions produced by humans when sweating and breathing. Through the consumption of food, nitrogen (N) is introduced in our system, and then disposed. Most nitrogen is released into the sewer system; the ammonia released through sweating and breathing is calculated within this emission source.

Contribution to the national emission

The contribution of this source to the total national NH_3 emission is around 1% (based on 2014).

19.2 Calculation

For the complete time series, the emissions are calculated as follows. Emission = Activity data x Emission factor

Activity data = the number of Dutch inhabitants Emission factor = kg emission per inhabitant

a) Activity data

The number of inhabitants in the Netherlands is derived from CBS Statline on an annual basis. The number of people living in the Netherlands at the end of June in a specific year is taken as activity data for that year.

b) Emission factor

With the food humans consume, nitrogen (N) is consumed. It is estimated that a human excretes 5 kg N (NH_3) per year in different ways (urine, sweat, faeces etc.) (*Battye et al 1994*). Most N or NH_3 is released with the urine and faeces, and it is assumed this goes through the sewer system.

The first emission factor used by the Dutch emission inventory was based on $van\ der\ Hoek\ (1994)$ who reported a total emission factor of 0.7 kg NH $_3$ per inhabitant per year, combining 0.3 kg NH $_3$ from sweating and breathing, use of ammonia as cleaning product (1 litre of ammonia solution per household) and the ammonia emissions of cats and dogs.

In another report by *Bouwman et al (1997)* an emission factor of the same magnitude is reported for human emission of NH_3 . In this study the emissions calculated are used for a global emission inventory. The author mentions that it is difficult to come to a good estimation of the emission factor, but describes that this source should not be neglected. Therefore, he assumes $0.5 \text{ kg } NH_3$ per person per year, independent of sanitary arrangements and includes domestic pets (cats/dogs). Since

the Dutch standard includes a good sewer system and the Netherlands reports the emissions of domestic pets separately, this emission factor is considered to too high for the Netherlands.

Joshua Fu et al (2010) report that, perspiration, respiration, untreated waste, cigarettes, household ammonia use, diapers and homeless people are sources of ammonia emissions directly caused by humans. They report an emission factor of 0.44 kg NH3 per person per year for all these emission sources. No separate emission factors are presented, though the distribution of the emission of the different sources is reported. Both perspiration and respiration are reported to contribute about 40% each. The emissions of untreated waste, household ammonia use, and homeless people contribute about 4-6% each. Cigarettes, (untreated) waste, and household ammonia are sources that are included as separate sources in the Dutch emission inventory. Other studies also report that the emissions from breathing are less than those from sweating. Therefore, the emissions in this document could be too high for the Netherlands.

Battye et al (1994) consider different references, varying from 0.25-1.3 kg NH₃/human/year from breathing and sweating. Although they mention that further research is needed, they recommend using the emission factor of 0.25 kg NH₃ p.p.p.y.; this emission factor was retrieved from a NAPAP report. The most interesting aspect is the reference to a measurement of NH₃ in a home. They report that an emission factor of 1 kg NH₃ should result in a concentration of about 431 μ g/m³, while the concentrations measured are between 32 and 39 μ g/m³. It could therefore be concluded, although this is not done in their study, that the emission factor is around 0.1 kg NH₃ per person per year.

One of the most comprehensive studies on the emissions of ammonia from non-agricultural sources was conducted by *Sutton et al (2000)*. In this report, the emissions of sweating are calculated with a range of emission factors from 2.08 g NH $_3$ till 74.88 g NH $_3$ (as g N per person and year). For breathing the range is 1.0-7.7 g NH3 (as N per person per year). *They* reference a number of reports and explain their assumptions. One of the most important assumptions made is the amount of NH $_3$ that volatilizes from sweat (10-30%). If no volatilisation is assumed, the highend emission factor is about 0.25 kg NH $_3$ per person per year. This is equal to that reported by *Battye et al (1994)* and a reference used by *Sutton et al (2000)*.

Furthermore, studies by *Chang (2014), Zheng et al (2012)* and *Klimont&Brink (2004)* use the emission factors presented by *Sutton et al (2000)*.

Some countries other than the Netherlands also report the emissions of human sweating and breathing, for example Switzerland, Canada and the UK (in the past). The three countries used the 'best' emission factors provided by *Sutton et al (2000) of 0.017 kg NH₃ p.p.p.y.* This is less than the ammonia emission factor noted in the *guidebook 2013* of 0.05 kg NH₃ per person per year.

Only one study (*Sutton et al 2000*) reports an emission factor for the ammonia emissions from diapers. Depending on the age and some assumptions, the emission factor ranges from $2.4-68 \text{ g NH}_3$ per infant

per year. A first estimate for children (age 0-3 year) in the Netherlands gives an emission of 2 to 50 tonnes NH_3 a year. Since this is only one reference and it forms a relatively low contribution to the national total, the decision was made not to include this emission (separately) in the Dutch emission inventory.

Emission factor used in the Netherlands emissions inventory

The high-end emission factors by *Sutton et al (2000)* are used in our calculations, resulting in a total emission factor of 0.0826 kg NH₃-N per person per year (sum of 74.88 and 7.7 gram p.p.p.y. for sweating and respiration respectively). For the Dutch national emission inventory, this was recalculated to 0.1004 kg NH3 per person per year. Because the emission factors in other reports are higher, it was decided to choose the *Sutton et al (2000)*_high-end emission factors, instead of the 'best' emission factors. This reduces the risk of underestimating the human ammonia emissions and emission sources not calculated (homeless people and diapers) can be neglected.

19.3 Uncertainty and Quality checks

The uncertainty in the number of inhabitants in the Netherlands is considered to be very small, therefore the uncertainty is qualified as A. The uncertainty in the emission factor is estimated to be relatively high as emission factors vary between different sources and the amount of ammonia volatilized is based on an assumption. Hence the uncertainty is qualified as D.

Quality codes

Substance	Activity data	Emission factor	Emission
NH ₃	Α	D	С

Quality checks

No sector-specific quality checks were performed. For the general QA/QC, see chapter 2.

19.4 Spatial allocation

The ammonia emissions of humans are spatially allocated in the Netherlands based on the inhabitants.

Emission source/process	Allocation-parameter	
Human ammonia emission; sweating	inhabitants	
and breathing		

Details available via

http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation)

19.5 References

Battye R. et al, 1994, Development and selection of ammonia emission factors, EC/R Incorporated, Durham, North Carolina 27707

Bouwman et al, 1997, A global high-resolution emission inventory for ammonia, global biochemical cycles, Vol. 11, No. 4, pages 561-587

- Chang Y.H., 2014, Non-agricultural ammonia emissions in urban China, Atmos. Chem. Phys. Discuss., 14, 8495–8531, retrieved 2016 from http://www.atmos-chem-phys-discuss.net/14/8495/2014/acpd-14-8495-2014.pdf Dutch Bureau of Statistics, https://www.atmos-chem-phys-discuss.net/14/8495/2014/acpd-14-8495-2014.pdf Dutch Bureau of Statistics, https://www.atmos-chem-phys-discuss.net/14/8495/2014/acpd-14-8495-2014.pdf Dutch Bureau of Statistics, https://www.atmos.chem-phys-discuss.net/14/8495/2014/acpd-14-8495-2014.pdf Dutch Bureau of Statistics, https://www.atmos.chem-phys-discuss.net/14/8495/2014/acpd-14-8495-2014.pdf Dutch Bureau of Statistics, https://www.atmos.chem-phys-discuss.net/14/8495/2014/acpd-14-8495-2014.pdf Dutch Bureau of Statistics, <a href="https://www.atmos.chem-phys-discuss.net/14/8495/2014/acpd-14-8495/2014
- EMEP/EEA emission inventory guidebook 2013, other sources and sinks activities 110701 110703
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- Joshua Fu et al, 2010, Quality Improvement for Ammonia Emission Inventory, Department of Civil and Environmental Engineering, University of Tennessee, Knoxville, TN 37996-2010
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19.6 Version, dates and sources

Version: 1.2 Date: May 2017

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20 Leather maintenance products and office supplies

This section describes the emissions of NMVOC from office supplies and leather maintenance products.

Process description	Emk_code	NFR_code	Target group
Leather maintenance products	0802800	2D3a	Consumers
Office supplies	0820600	2D3a	Consumers
Office supplies	0820601	2D3i	Trade and services

20.1 Description emission source

All emissions from leather maintenance products are ascribed to consumers. The emissions from office supplies are ascribed to consumers (50%) and trade and services (50%). Leather maintenance products consist of polishing wax, furniture polish, furniture cleaners, shoe polish, etc. Office supplies consist of tip-ex, ballpoints, fibre-tip pens, text markers, etc. All these products contain NMVOC which emits to the air on use.

Contribution to the national emission

The contribution to the national total is less than 1% for both the consumers and trade and services groups.

20.2 Calculation

The market share of VOC containing products was monitored by InfoMil on a regular basis in the KWS2000 project. They performed producer and supplier surveys. The surveys however, were performed a long time ago: leather maintenance products in 2000 and for office supplies in 1997 and surveys.

The NMVOC total is divided into the individual substances, using an average profile established by TNO and Association in 1992.

20.3 Quality codes

Substance	Activity data	Emission factors	Emission
NMVOC	D	Not relevant	D

20.4 Spatial allocation

Details are available at:

http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation)

The emissions NMVOC are allocated in the Netherlands based on:

Emission source/process	Allocation-parameter	Source data
Office supplies	Floor area commercial and industrial buildings	LISA
Office supplies	Population density	Bridgris (ACN)
Leather maintenance products	Population density	Bridgris (ACN)

20.5 References

KWS2000/InfoMil, 1999, Jaarverslagen 1996-2000, InfoMil, Den Haag.

20.6 Version, dates and sources

Version: 1.1

Date: November 2013

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21 Meat preparation and charcoal use (CRF 1.A.4.b)

This section describes emissions of charcoal use from barbecues and meat preparation.

Process description	Emk_code	CRF_code	NFR_code	Sector
Meat preparation	0801800	1A4bi	1A4bi	CON
Charcoal use	0801801	1A4bi	1A4bi	CON

21.1 Description emission source

NMVOC is produced during meat preparation as result of the heating and burning of grease. During charcoal burning on the barbecue, various compounds are emitted, for example particulate matter, CH_4 and NMVOC.

Contribution to the national emission

This emission is not a key source of greenhouse gases

The contribution of NMVOC emissions from meat preparation to the total national NMVOC emission was 0.5% in 2009.

The emission of CO_2 is booked as a memo item under the Kyoto protocol as charcoal is considered as biomass. The amount of CO_2 emissions caused by charcoal burning is negligible in relation to the national total of CO_2 emissions from burning biomass.

The contribution to national emissions of PM_{10} , $PM_{2.5}$, CO, CH_4 , NO_X , N_2O , SO_2 and NMVOC as a result of charcoal burning is negligible.

21.2 Calculation

Emissions are calculated as follows: Emission = Activity data x Emission factor

This is a tier 1 methodology. The methodology is consistent with the IPCC 2006 Guidelines.

a) Activity data

Since no actual data are available for the use of charcoal in Dutch households, all emissions are calculated based on the amount of meat used. Data on the number of inhabitants and the amount of meat consumed annually per inhabitant were gathered from the national statistics bureau. Combined, these result in a total amount of meat consumed in the Netherlands.

It has been estimated that 0.3% of the meat is prepared on the barbeque, and about 80% is fried, roasted or grilled.

Charcoal use is estimated by multiplying the amount of meat consumed by 0.194 TJ charcoal/kton meat. This is based on 1.5 kg charcoal per 1 kg meat prepared on the barbeque.

In 2009, this resulted in 1,430 kton meat being consumed and 278 TJ charcoal used on the barbecue

b) Emission factor

The emission factors for NMVOC, particulate matter, SO_2 , NO_X and CO are derived from Brouwer et al. (1994). The emission factor for CO_2 (memo item), CH_4 and N_2O are derived from the 2006 IPCC guidelines; default emission factors are used.

Substance	EF	Unit		
Meat preparation				
NMVOC	0.5	g/kg meat		
Barbecuing				
NMVOC	250	kg/TJ charcoal		
SO ₂	10	kg/TJ charcoal		
N ₂ O	1	kg/TJ charcoal		
NO _X	50	kg/TJ charcoal		
CO	6000	kg/TJ charcoal		
CH ₄	200	kg/TJ charcoal		
PM ₁₀	150	kg/TJ charcoal		
PM _{2.5}	75	kg/TJ charcoal		
CO ₂ (memo item)	112	ton/TJ charcoal		

21.3 Uncertainty and Quality checks

The activity data for burning charcoal in households are estimated based on the amount of meat consumed yearly. This in turn is based on the assumption that barbecuing is solely responsible for charcoal usage in households. Since the amount of charcoal used in the Dutch households is based on meat consumption combined with estimated charcoal sales, the uncertainty is estimated at 50%, based on expert judgement. The emission factors (and corresponding uncertainties) used for charcoal burning are derived from 'IPCC guidelines 2006'. Therefore, the uncertainty bandwidth for N_2O ranges from -62.5% to 275%. For CH_4 , the uncertainty bandwidth is -66.6% to 200%. The corresponding uncertainty of CO_2 (memo-item) is reported as 20%.

The other emission factors are estimated based on a single report from the US; these emission factors are therefore not very reliable; The reliability of the data is qualitatively indicated in the table below with codes A-E (see Appendix A). The valuations are based on expert judgement.

Quality codes

Substance	Activity data	Emission factor	Emission
NMVOC (meat)	В	С	С
NMVOC (BBQ)	В	С	С
SO ₂	В	С	С
N_2O	В	В	В
NO_X	В	С	С
CO	В	С	С
CH ₄	В	В	В
PM	В	С	С
CO ₂	В	В	В

Quality checks

No sector-specific quality checks are performed. For the general QA/QC, see chapter 2.

21.4 Spatial allocation

Consumer emissions are spatially allocated in the Netherlands based on population density.

Emission source/process	Allocation-parameter	
Meat preparation	Population density	
Charcoal use in barbecue	Population density	

Details available via:

http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation)

21.5 References

CBS statline

Brouwer J.G.H., J.H.J. Hulskotte, C.H.A. Quarles van Ufford, 1994, Vleesbereiding, inclusief gebruik barbecue, WESP-rapport C-2, RIVM-rapportnr 773009003.

ER, 2009. Netherlands Emission Registry. Data on 2009 available from www.emissieregistratie.nl

IPCC 2006 guidelines. http://www.ipcc-nggip.iges.or.jp/public/index.html

21.6 Version, dates and sources

Version: 1.2 Date: May 2015

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22 Paint

This section describes the non-methane volatile organic compounds (NMVOC) emission from paint.

Process description	Emkcode	NFR code	Sector
Car paint – use of paint and lacquer	8920800	2D3d	Trade and services
Paint - construction	0802200	2D3d	Construction
Paint - consumers	0802201	2D3d	Consumers
Road paint	0119800/ 0129800	2D3d	Construction

22.1 Description emission source

Paint may contain NMVOC which evaporates to the air during and after use. Paint includes products like coating, (wall) paint, lacquer, varnish, plaster, glue, stripper and filler and thinner. The Netherlands Association of Paint Producers (Vereniging Van Verf en drukinkt Fabrikanten (VVVF)) provides annual sales data for the calculation of NMVOC emission.

Contribution to the national emission and sector totals

An overview of the contribution of NMVOC by paints to the target group and the national total can be found in the following table.

Sub market VVVF	Target group	Contribution to Target group (%)	Contribution to national total (%)
Car repair	Trade and	10	1.3
lacquer	services		
Construction	Construction	94	3.7
Consumers	Consumers	7	1.5

22.2 Calculation

The annual national paint sales, including information on NMVOC content, are provided by the VVVF annual paint sales statistics, representing about 95% of the Dutch total market. The remaining 5% consists of directly imported paint. The VVVF divides different sub markets as shown in the table below. The ENINA task force covers the industry target group and therefore this is not addressed in this report.

Sub market VVVF	Target group
Car repair lacquer	Trade and services
Construction (including steel	Construction
preservation and road marking)	
Consumers	Consumers
Industry and carpentry factories	Industry
Ship building	Industry

Car repair lacquer

The total NMVOC emission from this source is calculated as follows:

EM_{totalcrl} = EmN + EmI EmN = NMVOC content national paint sales EmI = NMVOC content directly imported paint

It is assumed that:

- all paint sold will be used the same year and that the NMVOC emitted is 100% and the imported paint has the same NMVOC percentage as the paint sold by VVVF;
- 5% is directly imported paint.

Construction (including steel preservation and road marking)

The total NMVOC emission from this source is calculated as follows:

Emtotalconstruction = EmN-C + EmI-C + EmN-SP + EmI-SP
EmN-C = NMVOC content national paint sales of construction
EmI-C = NMVOC content directly imported paint of construction
EmN-SP = NMVOC content national paint sales of steel preservation
EmI-SP = NMVOC content directly imported paint of steel preservation

It is assumed that:

- all paint sold will be used the same year and that the NMVOC emitted is 100%;
- the NMVOC percentage of the imported construction paint is 2%;
- the imported paint of steel preservation has the same NMVOC percentage as the paint sold by VVVF;
- for construction, 35% is directly imported paint, and for steel perservation, 10% is directly imported paint.

The total NMVOC emission from the construction sector is divided into road markings and others, based on the amount of road markings.

Consumers

The total NMVOC emission from this source is calculated as follows:

EM_{totalconsumers} = EmN + EmI EmN = NMVOC content of national paint sales EmI = NMVOC content of directly imported paint

It is assumed that:

- all paint sold will be used the same year and that the NMVOC emitted is 100% and the imported paint contains the same amount of NMVOCs as the paint sold by VVVF,
- 0% is directly imported paint.

a) Activity data

Total NMVOC emission is subdivided into individual substances based on paint profile statistics provided by the VVVF (VVVF 1997).

Substance in paint profile	Factor*
Additional Nonhalogenated volatile hydrocarbons	0,119
Additional Alif nonhalogenated hydrocarbons	0,264
Additional Aromatic nonhalogenated hydrocarbons	0,045
Methylenechloride	0,004
Ethanol	0,015
Esters boiling point <150°C	0,224
Ketone	0,075
Propyleneglycomethylether	0,045
Propyleneglycomethylether acet	0,045
Toluene	0,030
Xylene	0,134

^{*}Based on VVVF statistics 1997

22.3 Uncertainty

The uncertainties of the emission calculation were quantified by Utrecht University (J. vd Sluys) in 2002.

Quality codes

Substance	Activity data	Emission factor	Emission
NMVOC			С

22.4 Spatial allocation

The emissions of consumers and trade and services are allocated in the Netherlands based on population density. The emission of road paint is allocated based on road density.

22.5 References

Instituut voor toegepaste milieu-economie (TME)

2003. Kosteneffectiviteit VOS maatregelen 2010. eindrapportage.

Instituut voor toegepaste milieu-economie (TME) 2003.

Kosteneffectiviteit VOS maatregelen 2010. achtergronddocument verf, bouw en doe het zelf.

Ministerie van VROM 2005. Nationaal Reductieplan NMVOS industrie, HDO en bouw, bijdrage van de sectoren aan het realiseren van het NEC-plafond in 2010.

Sluijs vd J., et al., 2002. Uncertainty assessment of VOC emissions from paint in the Netherlands. Utrecht University, NW&S E-2002-13.

Staatsblad 2005/632. Besluit organische oplosmiddelen in verven en vernissen (BOOVV).

Vereniging Van Verf en drukinktFabrikanten (VVVF) 1999. Grondstoffenverbruik in 1997 in de Nederlandse verfindustrie.

Vereniging Van Verf en drukinktFabrikanten (VVVF) 2002.VOSreductieplan 2010 voor de verf-en drukinktindustrie.

VVVF Annual reports on www.vvvf.nl

22.6 Version, dates and sources

Version: 1.1

Date: March 2015

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23 PCP pressure treated wood

This section describes the emissions of PCP and dioxins from resident façade boarding treated with PCP. The section does not describe the NMVOC emission.

Process description	Emk_code	NFR_code	Target group
PCP pressure treated wood	0010300	2D3i	Consumers

23.1 Description emission source

In the past, wooden façade boarding of residences was treated with PCP (pentachlorophenol). Over time, PCP and dioxins are emitted from the wood to the air. Dioxin emissions occur because wood was treated with contaminated paint.

Contribution to the national emission

PCP from façade boarding is the only source within the ER. The contribution of dioxin is ~46% of the national total.

23.2 Calculation

Dioxin

In 1990, Bremmer at al. estimated the dioxin emission from façade boarding in the RIVM report 'emissies van dioxines in Nederland' (Bremmer et al 1993). The use of PCP was prohibited in 1989 and therefore it is assumed that there has been a linear decrease of emissions. Bremmer et al. estimated emission of $\sim\!25$ g I-TEQ (1990), $\sim\!20$ g I-TEQ (2000) and $\sim\!10$ g I-TEQ in 2020.

Pentachlorophenol

In a PCP base document, Slooff et al. (1990) calculated an emission of 35 tons in 1987. The PCP emission was set to 34 tons for the year 1990. According to Slooff et al and Bremmer et al. the emission will be reduced to 50%, therefore, an annual emission of 3.3% is assumed.

The emissions of PCP and dioxin are also reduced because façade boarding has been replaced. Assumptions for reductions in the total amount of façade boarding are 1% each year between years 10 to 20, 2% for each year between year 20 to 30, 3% for each year between years 30 to 40, and 4% for each year after year 4.

The emission values for PCP and dioxin presented by Slooff et al and Bremmer et al include the reduction of the amount of façade boarding reduction after 10 years.

Measures influencing the calculation The use of PCP was prohibited in 1989.

23.3 Uncertainty

The uncertainties of the emission calculation are quantified in the reports by Bremmer et al and Slooff et al.

Bremmer calculated an average of 16 g I-TEQ dioxin in 1990 and stated in the report that the maximum could not exceed 25 g I-TEQ dioxin. For 1990, the ER assumed an emission of 25 g I-TEQ. This would mean that there uncertainty is mainly in the low end of the value. The 95% confidence interval is skewed. As a rule in the ER, the highest uncertainty value is used and the 95% confidence interval is normally distributed around the 25 g I-TEQ.

Slooff et al reported the average PCP emission. The 95% confidence interval is normally distributed around the average.

Quality codes

Substance	Activity data	Emission factors	Emission
Dioxins	D	Not relevant	D
PCP	D	Not relevant	E

23.4 Spatial allocation

The dioxin and PCP emissions are allocated in the Netherlands based on population density. Details are available at:

http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation)

23.5 References

Bremmer, H.J., L.M. Troost, G. Kuipers, J. de Koning, A.A. Sein, 1993, Emissies van dioxinen in Nederland, RIVM rapportnummer 770501003.

Slooff et.al, Basisdocument PCP, RIVM, november 1990.

23.6 Version, dates and sources

Version: 1.1

Date: November 2013

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24 Petrol stations

This section describes the NMVOC emissions from petrol stations.

Process description	Emk_code	NFR_code	Target group
NACE 47.3: gas stations, spills tank refill	8920900	1B2aiv	Trade and service
NACE 47.3: gas stations, vapour expel - tank refill	8920901	1B2aiv	Trade and services
NACE 47.3: gas stations, vapour expel - storage tanks	8920902	1B2aiv	Trade and services

24.1 Description emission source

The term 'Petrol stations' includes the distributions points for road traffic as well as petrol stations on company grounds (meant for company cars). The NMVOC emissions of petrol and LPG are reported. LPG is reported as butane and propane (50/50). Emission occurs during the filling of tanks and results from two sources: the loss due to leakages of the fuel (petrol) and loss due to expulsion while filling car tanks and storage tanks (petrol and LPG). In the Netherlands, there are between 40 and 45 filling stations for cars using natural gas as fuel. However, it is assumed that the losses occurring during filling of natural gas are negligible. Before disconnecting the filling pistol, the natural gas in the dead space between pistol and tank is recovered; no gas is emitted into the air.

Contribution to the national emission

The contribution of NMVOC from petrol station is around 5.5% of the national total NMVOC emission and approximately 18% of the target group trade and services (ER, 2009).

24.2 Calculation

Emissions are calculated as follows: Emission = Activity data x Emission factor

Activity data = amount of fuel used in the Netherlands for road transportation

Emission factor = emission per litre fuel used

a) Activity data Leakage losses

During filling car tanks spillage of petrol can occur.

On average, it is assumed that the minimal spillage is at least 1-2 ml and the average tank amount per filling is 40 litres. The density of petrol is 0.72 kg/litre. It is therefore assumed that a million litres of petrol produce 720 tons of VOC.

In the last decade, the Netherlands used 5.5 billion litres petrol for transportation by road (statline.CBS.nl). Based on these assumptions and official data, we calculated a VOC emission of around 300,000 kg.

Since the amount of petrol has been constant over time, the VOC emission due petrol spillage has not changed.

After disconnecting the LPG filling pistol, LPG (a mixture of butane and propane 50/50) will be-emitted into the air. The average dead volume of the pistol and connection nipple of the tank is 12.5 ml (personal communication LPG installation branch).

It is assumed that, on average, the tank is filled with 40 litres LPG. CBS (statline.CBS.nl) in the Netherlands provides data for the amounts of LPG used for transportation by road. Based on these assumptions and the official data, we calculated an LPG spillage of 97 tons.

Expulsion losses car tanks

At the start of refuelling with petrol, the tank is filled with petrol vapour. When petrol flows in the tank, the petrol vapour is emitted (Bernouille-principal). Therefore, during refuelling, petrol is emitted into the air. In the Netherlands, measures were implemented to reduce the emission of petrol. These stage 1 and stage 2 measures were implemented in 2000 and 2005 respectively. From 2005, the expulsion losses have been settled at 1.27 kt. Because the amount of petrol for transportation has not changed (5.5 billion litres according to statline.CBS.nl) the expulsion losses have been constant for years.

Measures influencing the calculation

Although both the stage 1 and the stage 2 measures have been implemented since 2005, a rest emission of 25% of pre-implementation emission will remain. No further emission reduction measures are foreseen. Although mobility is increasing, the sale of petrol remains constant. The sale of diesel however has increased.

b) Emission factors

The factors for the emission by petrol leakage are unknown.

NMVOC (kT)	Amount Petrol	Expulsion losses		Spillage	Total	Realisation
	(billion			refueling		Stage I and
year	litre)	storage	car's	cars	losses	H
1980	?	5.1	4.9	0.6	10.6	
1990	?	4.9	4.9	0.6	10.4	
2001	5.5	0.0	1.9	0.4	2.5	Stage 1
2005	5.5	0.0	1.3	0.3	1.6	Stage II
2011	5.7	0.0	1.3	0.3	1.6	
2012	5.4	0.0	1.3	0.3	1.6	

Substance in LPG	Emission factor
propane	0.5
butane	0.5

Butane/propane (kT)	Amount Petrol	Losses	Spillage r	efueling c	ars (kT)
	(billion			wv	WV
year	litres)	storage	LPG total	butane	propane
1985	1.5	0.0	0.26	0.13	0.13
1990	1.5	0.0	0.24	0.12	0.12
2001	1.0	0.0	0.16	0.08	0.08
2005	0.7	0.0	0.11	0.06	0.06
2011	0.5	0.0	0.09	0.045	0.045

24.3 Uncertainty

The uncertainties of the emission calculation are not quantified.

Quality codes

Substance	Activity data	Emission factor	Emission
NMVOC	В	С	С

24.4 Spatial allocation

The emissions of the petrol stations are assigned to the locations of the petrol stations (SBI 47.3) in the Netherlands according the ratio of employees at the petrol stations. Details are available at: http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation).

24.5 References

CBS, Statline www.statline.CBS.nl

Comprimo (briefrapport 18 november1994) over schattingen voor lekverliezen van benzine.

ER, 2009. Netherlands Emission Registry. Data on 2009 available www.emissieregistratie.nl

VROM, Besluit 'tankstations en milieubeheer' en 'herstelinrichtingen voor motorvoertuigen en milieubeheer' en het wijzigingsbesluit daarop (Stb 1996, 228).

24.6 Version, dates and sources

Version: 1.1

Date: November 2013

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25 Residential combustion, Wood stoves and Fireplaces (CRF 1.A.4.b)

This section describes the emissions from wood combustion by consumers.

Process description	Emk_code	CRF_code	NFR_code	Sector
Residential wood	T012200	1A4bi	1A4bi	CON
combustion				

25.1 Description emission source

In the Netherlands, residential combustion of wood is mainly for creating a homely ambiance. Although wood combustion in stoves is sometimes considered more environmentally friendly, hardly any homes in the Netherlands use wood combustion as their main heat source.

Wood combustion in stoves and fireplaces leads to various emissions. The emissions of CO_2 are the result of biomass burning and therefore reported as such. In addition to CO_2 , CH_4 , N_2O , NO_X , particulate matter, VOCs and other compounds are emitted.

Contribution to the national emission

This emission is not a key source of greenhouse gases.

Wood stoves and fireplaces are relatively the main contributors to emissions of particulate matter accounting for about 6% of the national total; this is particulate matter without condensables. If condensables are included in the particulate matter emissions, the contribution increases to about 14% of the national total.

All contributions to the national emissions are related to the emissions in the year 2009.

25.2 Calculation

The emissions from fireplaces and different types of woodstoves are calculated using a model described in Jansen (2010) and Jansen (2014).

The emissions are calculated as follows for each type of stove; Emission = Activity data x Emission factor

For the greenhouse gases this is a tier 1 methodology as there is no difference in the emission factors. The methodology is consistent with the IPCC 2006 Guidelines.

a) Activity data

For the year 2009, this results in a total wood use of about 1200 million kilograms in all stoves and fireplaces. This is the equivalent of 17000 TJ wood burned in the stoves and fireplaces. The amount of wood is based on Segers (2013) and on earlier studies on wood burning in stoves for

the Netherlands, and modelled in the emission model (for more details see Jansen 2010).

b) Emission factor

All emission factors are reported in the 2010 and 2014 reports about the emission model for woodstoves (Jansen 2010; 2016).

The emission factors for some substances are listed in the following table.

	EF (type of stove)				
Substance	Fire place	conventional	approved	DINplus	unit
CO ₂	112	112	112	112	kg/GJ
CH ₄	0.3	0.3	0.3	0.3	kg/GJ
N ₂ O	4	4	4	4	g/GJ
PM ₁₀	161	194	97	52	g/GJ
Condensables	484	323	129	80	g/GJ

The default emission factors from the IPCC 2006 guidelines are used for greenhouse gas emission.

25.3 Uncertainty and Quality checks

The activity data for wood burning is calculated yearly, based on 5-yearly questionnaires. The uncertainty might therefore fluctuate over time and is estimated at 35% (ND) based on expert judgement in combination with the data in *Segers*, 2010.

The emission factors for greenhouse gases are based on the 'IPCC guidelines 2006'; the corresponding uncertainties are reported as: 20% for CO_2 (reported as memo-item), -62.5% to 275% for N_2O , and -66.6% to 200% for CH_4 .

For the other substances, the uncertainty in the emission factors has not been specifically determined. Based on expert judgment, the uncertainty is rated as C (see Appendix A).

Quality checks

No sector-specific quality checks are performed. For the general QA/QC, see chapter 2.

25.4 Spatial allocation

The emissions of consumers are spatially allocated in the Netherlands based on inhabitant distribution.

Emission source/process	Allocation-parameter
Burning wood in Stoves	Population density

Details available via

http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation)

25.5 References

Jansen B.I., 2010, Emissiemodel houtkachels, TNO, The Netherlands Jansen B.I., 2016, Emissiemodel houtkachels update, TNO, The Netherlands

Oldenburger et al., 2012, Houtstromenstudie, Probos, The Netherlands Segers R., 2010, Houtverbruik huishoudens, CBS, The Netherlands Segers R., 2013, Houtverbruik huishoudens Woon 2012, CBS, The Netherlands

25.6 Version, dates and sources

Version: 1.2 Date: March 2016

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B.I. Jansen	TNO	bart.jansen@tno.nl

26 Shooting

This section describes the emissions of lead to the soil from shooting.

Process description	Emk_code	Target group
Shooting	E800200	Trade and services

26.1 Description emission source

The emission source is related to clay-pigeon shooting. Hunting and traditional shooting is not included. Since 2008, the use of lead has been prohibited in both disciplines and has been replaced by steel. The emission of steel (iron) is reported by the agriculture group. Only official clay-pigeon shooters (match) had an exemption from the minister until 2016 to shoot with lead, but this may be continued because lead is still used in international competition. Clay-pigeon shooting with lead is only performed at one location in the Netherlands, in Emmen. Currently the number off match shooters is small.

Contribution to the national emission

The lead emission contribution from clay-pigeon shooting to the national total is $\sim 3\%$. (ER 2012).

26.2 Calculation

The use of lead-shot has not been permitted since December 2004. According to the information from the KNSA (Koninklijke Nederlandse Schutters Associatie) there are currently only 3 competitive clay-pigeon shooters. Official clay-pigeon shooters (match) were permitted to use lead by the minister until 2016. A competitive shooter trains on a regular basis and shoots 14000 times per year. A shell contains 24 grams lead. Thus, a shooter uses 336 kg lead. In 2012, there were 3 competitive shooters in the Netherlands, thus, the emission of lead due to clay-pigeon shooting was ~1 ton. According to the KNSA, the lead is not removed from the shooting range.

26.3 Quality codes

Substance	Activity data	Emission factors	Emission
Lead	В	Α	С

26.4 Spatial allocation

There is only one location in the Netherlands, Schietsportcentrum Emmen in Emmer-Compascuum.

26.5 References

Bon, 1988, Bon, J. van en J.J. Boersema, feb 1988, Metallisch lood bij de jacht, de schietsport en de sportvisserij, IVEM rapport nr 24, Groningen.

- Booij, 1993, Booij, H ,et al, sept 1993, Alternatieven onder schot, RIVM rapport nr 710401026, RIVM, Bilthoven.
- De straat, 1996, De Straat Milieu-adviseurs, 1996, Beperking van de milieubelasting bij kleiduivenschieten, Projectnummer B2112, De Straat Milieu-adviseurs, Delft.
- KNSA, 2011/2013, Telefonisch onderhoud met Dhr. Duisterhof van de KNSA. http://www.knsa.nl
- Staatscourant 2004/237, Besluit kleiduivenschieten WMS, 19 mei 2004
- Staatscourant 2013, Beschikking van 11 juni 2013, Ontheffing verbod op gebruik van loodhagel bij het kleiduivenschieten voor topsporters. Staatscourant nr 17795
- VROM, 1999, Traditioneel Schieten, Circulaire, VROM/DGM/SVS, 1 november 1999
- VROM, 1995, VROM/DGM, dir. Stoffen, Veiligheid, Straling, afd. Stoffen, 1995, Circulaire Beperking loodbelasting van de bodem bij traditioneel schieten, VROM/DGM, Den Haag.

26.6 Version, dates and sources

Version: 1.1

Date: November 2013

Administrator	Organisation	E-mail address
J. Meesters	RIVM	Joris.Meesters@rivm.nl

27 Wholesale Business in fuel and remaining mineral oil products

This section describes NMVOC emissions from Wholesale business in fuel and remaining mineral oil products. Prior to 2010, this emission source was addressed using the terms "petrol distribution points" or petrol distribution chain".

Process description	Emk_code	NFR code	Sector
Wholesale business in fuel and	8921100	1B2aiv	Trade and
remaining mineral oil products.			services

27.1 Description emission source

The petrol distribution points are viewed as the link between petrol stations and the refinery. In the KWS2000 project this was termed "petrol distribution chain". The emissions resulting from loading tanklorries at refineries and filling large storage tanks are incorporated in the emission trend of the distribution chain. Since 2001, the expulsion losses when filling tank-lorries and storage tanks have no longer been reported individually, but as a total and imply emissions resulting from petrol only.

Contribution to the national emission

The NMVOC emissions from the petrol distribution chain decreased after a light increase in 1996. The decrease was mainly due to reduced NMVOC emission from storage tanks.

The contribution of NMVOC from the wholesale business and remaining mineral oil products is <1% of the total national NMVOC emission and approximately 7% of the NMVOC emission from HDO (ER, 2009).

27.2 Calculation

Until 1999, emission data was based on data for 1997 and overall estimates of branch developments in the following years. In 2001, the petrol storage facilities were sent questionnaires in order to gain information about the implementation of measures and the residual emissions in 2000. As the response to this questionnaire was limited, it was not possible to use this questionnaire in order to obtain a good picture of the VOC emissions and the degree of implementation. The individual companies have therefore not gained any insights in their VOC emissions. With regard to the degree of implementation of measures, the questionnaire data show that the major petrol storage facilities have implemented these measures, but the situation at the smaller storage facilities remains unclear (KWS 2000, 2002).

In 2004, the VPNI (Vereniging Nederlandse Petroleum Industrie: Association of the Dutch Petrol Industry) developed a reduction plan for VOC emissions for the years 2000-2010. The VPNI produced emission calculations based on measures mentioned in KWS 2000. *Measure affecting the calculation*

Due to the KWS2000, the following precautions were undertaken for the distribution chain:

- storage tanks with internal floating deck;
- · treatment of expulsion air.

These measures were established in the departmental regulation "Storage, transfer and distribution of petrol environmental management" in December 1995. No further reductions were presented in the reduction plan VOS 2000-2010 for petrol distribution (VPNI). Since 2003 the emissions from distribution points are kept constant.

27.3 Uncertainty

The uncertainties of the emission calculation are not quantified. Two general measures were taken (labelling and environmental car), however the effects of these measures are unknown.

Quality codes

Substance	Activity data	Emission factor	Emission
NMVOC	С	С	С

27.4 Spatial allocation

The emissions of the Wholesale Business in fuel and remaining mineral oil products are assigned to the locations of the petrol distribution points (SBI 46.71) in the Netherlands according to the ratio of employees at the distribution point. Details are available at:

http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation).

27.5 References

ER, 2009. Netherlands Emission Registry. Data on 2009 available from www.emissieregistratie.nl

VNPI, P.Houtman, mrt 2004 Reductieplan VOS 2000-2010 voor de aardolieketen.

VROM Ministeriele regeling Op-, overslag en distributie benzine milieubeheer, 27 december 1995.

27.6 Version, dates and sources

Version: 1.1

Date: November 2013

Responsibility: manager WESP task group

Administrator	Organisation	E-mail address	
J. Meesters	RIVM	Joris.Meesters@rivm.nl	

Note: Since 1999, no new data is available on emission variables and emission factors. Therefore, the data might be outdated.

28 Smoking of cigarettes and cigars

This section describes the emissions caused by smoking cigars and cigarettes.

Process description	Emk_code	NFR_code	Sector
Smoking of cigars	0801001	2D3i	CON
Smoking of cigarettes	0801002	2D3i	CON

28.1 Description emission source

When tobacco products are burned, the fumes contain a mix of different substances. These substances are inhaled and exhaled by the smoker or emitted to the air directly. Since smoking has been found to be unhealthy, the number of people smoking has declined almost yearly, with a comparable reduction in the consumption of tobacco products.

Contribution to the national emission

The contribution of particulate matter for smoking to the total national PM10 emission is <5%.

For other important (PRTR) substances the contribution to the national total emission is < 0.5%.

28.2 Calculation

Emissions are calculated as follows: Emission = Activity data x Emission factor

Activity data = Number of cigars (or cigarettes) smoked Emission factor = kg emission per number of cigars (or cigarettes) smoked

a) Activity data

Cigarettes

For the number of cigarettes, most data are provided by the branch organisation for the cigarette industry (Stichting Sigarettenindustrie). The data cover the years 2005-2014 for the number of cigarettes only. For rolling tobacco, information is retrieved from the branch organisation for rolling tobacco 'VNK' (Vereniging Nederlandse Kerftabakindustrie). For rolling tobacco, the VNK presents information for the years 2003-2013.

For the years 1990-2010, the Dutch Centre for Statistics (CBS) provided information for the total amount of rolling tobacco and cigarettes. All of the above data match for the years covered by multiple data sources, however it seems that the data is not independent but comes from the same source.

In recent years, starting in 2014, data are provided by the Trimbos institute, a research organisation on mental health and addiction.

Cigars

For the years 1990 to 2006, the Dutch Centre for Statistics (CBS) presented information on the consumption of cigars in the Netherlands. This information was the average number of cigars smoked per inhabitant. However, from 2006 this information is no longer available from the CBS, so in 2014 contact was made with the Dutch Cigar Manufacturers Association (NVS). The NVS was willing to provide data annually starting with the year 2013. The number of cigars consumed in the years between 2006 and 2013 are interpolated.

b) Emission factor

The emission factors for smoking are based on Brouwer et al 1994, written specifically for the WESP taskforce. For cigars, the emission factors are based on a report from the American National Cancer Institute.

Cultatanas	Cigarettes		Cigars		
Substance	EF	Unit	EF	Unit	
Acetaldehyde	4507	kg/ milliard pieces	2	mg/ piece	
NOx	2552	kg/ milliard pieces	0.5	mg/ piece	
Hydrogen cyanide	199	kg/ milliard pieces	2	mg/ piece	
NH_3	9478	kg/ milliard pieces	0.07	mg/ piece	
CO (till 2001)	64389	kg/ milliard pieces	119	mg/ piece	
Starting 2002	10000		161		
CO ₂ (memo)	294242	kg/ milliard pieces	305	mg/ piece	
PM ₁₀	52293	kg/ milliard pieces	63	mg/ piece	
PM _{2.5}	52293	kg/ milliard pieces	63	mg/ piece	
NMVOC	22767	kg/ milliard pieces	Not repo	rted	
Acetone	663	kg/ milliard pieces	Not repo	Not reported	
Acrylonitrile	57	kg/ milliard pieces	Not repo	rted	
Acrolein	Not reported		0.13	mg/ piece	
Acetic acid	Not reported		0.70	mg/ piece	
Formic acid	Not reported		0.26	mg/ piece	
Benzene	Not reported		0.39	mg/ piece	
Phenol	Not reported		0.15	mg/ piece	
Pyridine	Not reported		0.23	mg/ piece	
Anthracene	Not reported		0.27	µg/ piece	
Fluoranthene	Not reported		0.46	µg/ piece	
Pyrene	Not reported		0.40	µg/ piece	
Benzo(a)anthracene	Not reported		0.15	µg/ piece	
Benzo(a)pyrene	Not reported		0.09	µg/ piece	
Cadmium	Not reported		0.05	µg/ piece	
Copper	Not reported		0.23	µg/ piece	
Lead	Not reported		0.50	µg/ piece	
Nickel	Not reported		11	µg/ piece	
Zinc	Not reported		3.3	µg/ piece	

28.3 Uncertainty

The activity data is reported by reliable third parties. However, due to inconsistency between years, the uncertainty with respect to illegal imported tobacco products, and the data from branch organisations, the activity data is rated with an B.

The emission factors are based on old reports; for cigars and cigarettes an average weight is assumed. Therefore, the emission factors are rated with a C.

Overall, the emissions are rated with an C.

Substance	Activity data	Emission factors	Emission
All	В	С	С

28.4 Spatial allocation

The emissions of consumers are regionalised in the Netherlands based on population density.

Source	Allocation-parameter
Smoking	Population density

Details are available via:

http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation)

28.5 Reference

Brouwer J.G.H., J.H.J. Hulskotte, H. Booij, 1994, Roken van tabaksproducten, WESP Rapportnr. C4, RIVM-rapportnr. 773009006.

28.6 Version, dates and sources

Version 1.1

Date: February 2018

Administrator	Organisation	E-mail address
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29 Service stations, anti-corrosive treatment

This section describes the emissions caused by the anti-corrosive treatment of cars.

Process description	Emk_code	NFR_code	Sector
Service stations, anti-	8920700	2D3i	Trade and
corrosive treatment			Services

29.1 Description emission source

When old cars lose their corrosive protection, they are treated with anticorrosive products. This is mainly true for older cars because newly produced cars are generally equipped with a chassis of anti-corrosive materials such as plastics or galvanized steel. This treatment is mainly done by specialised service stations (garages). Due to decreasing demand, the number of workshops providing this service is decreasing. The anti-corrosive products used are a source of NMVOC, the NMVOC emissions are released in a short period after the product is applied.

Contribution to the national emission

The contribution of NMVOC from anti-corrosive treatment to the total national NMVOC emission is <1%.

29.2 Calculation

Between 1990 and 1999, the sale of anti-corrosive products was estimated by the company with the largest market share of the anti-corrosive product (KWS2000).

Starting with the year 2000, the emissions are calculated as follows: Emission = Activity data x Emission factor

Activity data = Number of cars treated Emission factor = kg emission per treated car

a) Activity data

It is assumed that only cars produced in 1985 or earlier need an anticorrosive treatment. The number of old cars in the Netherlands is based on data from Statistics Netherlands (CBS). They provide the number of cars still in use in the Netherlands which were manufactured before 1986. Not all those cars have to be treated each year. Under normal circumstances the treatment is only needed once every ten years. However, it is thought that most old-timers are treated more frequently, hence it is assumed that these cars are treated each eight year, or that 12.5% of the old-timers are treated.

b) Emission factor

For each car treated, the emission of NMVOC is estimated at 8 kilograms. This is based on the EMEP/EEA air pollutant emission inventory guidebook (2.D.3.d. table 3-6). However, this emission factor is not specific for this type of treatment, but for a general coating application.

The emission of NMVOC is distributed to individual substances based on an emission profile created by TNO in 1992.

29.3 Uncertainty

Up until 1999, the amount of NMVOC emission was estimated by expert judgement. This results in an uncertainty qualified with an C. The total number of cars is reported by a reliable third party starting from the year 2000. However, the number of cars treated is based on literature study combined with expert judgement. Therefore, the activity data should be rated with a B.

The emission factor for NMVOC is based on the EMEP/EEA guidebook, with the concession that the emission factor is not specific for this type of treatment. Therefore, the emission factor is rated with a D. Overall the emissions for the whole time-series are rated with a C.

Substance	Activity data	Emission factors	Emission
NMVOC	В	D	С

29.4 Spatial allocation

The emissions of consumers are regionalized in the Netherlands based on the amounts of employees within the service stations (SBI 50: garages).

Source	Allocation-parameter
Anti-corrosive treatment	Employee density within SBI 50

Details are available via:

http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation)

29.5 Reference

CBS statline. https://opendata.cbs.nl/statline/#/CBS/nl/

Dutch Emission Inventory.

http://emissieregistratie.nl/erpubliek/bumper.en.aspx

European Environment Agency.

https://www.eea.europa.eu/publications/emep-eea-guidebook-2016

KWS2000, multiple years, Annual reports. InfoMil, Den Hague.

29.6 Version, dates and sources

Version 1.1

Date: February 2018

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30 Manure from Domestic animals

This section describes the emissions of NH3 caused by domestic animals.

Process description	Emk_code	NFR_code	Sector
Manure of domestic animals	0802000	6A	Consumers

30.1 Description emission source

This emission source is calculated using the emissions from domestic animals. Domestic animals are defined as animals not used as livestock in the agricultural industry, with the exception of horses and ponies, which are calculated by the agriculture and nature task force. When animals consume food, the nitrogen (N) from the food is (partly) rereleased. Most N is released through the excretion of faeces and urine, which results in the emission of ammonia. Emissions of other substances caused by domestic animals are considered irrelevant and therefore are not calculated.

Contribution to the national emission

The contribution of this source to the total national NH3 emission is >5%.

30.2 Calculation

For the complete time series, the emissions are calculated as follows: Emission = Activity data x Emission factor

Activity data = Amount of households Emission factor = kg NH3 per household

a) Activity data

The number of households is derived from Statistics Netherlands.

b) Emission factor

The emission factor used by the Dutch emission inventory is based on Booij 1995, who calculated a total emission of 1220 tons of NH3 from domestic animals (cats, dogs, rabbits and birds) for the year 1990. The emission factors for cats and dogs calculated by Booij 1995 are 0.18 and 0.36 kg NH3 per animal and per year respectively. The emission calculated for cats and dogs by Booij 1995 is about 70% of the total NH3 emission from pets. With the total emission in 1990 and the number of households in 1990, an emission factor of 0.2 kg NH3 per household was calculated.

Some other authors, Joshua Fu et al 2010 and Bouwman et al 1997, report emission factors of around 0.7 kg NH3 per year for cats and around 2 kg NH3 per year for dogs. This is high when compared to Booij 1995. Furthermore, most other reports seem to base their emission factor on the work of Sutton et al (2000). The emission factor presented

in Sutton et al (2000) is 0.61 kg NH3 for dogs and 0.11 kg NH3 for cats; each per animal and per year.

30.3 Uncertainty and Quality checks

The number of cats and dogs is based on a survey commissioned by DIBEVO. A sample survey generally has a relatively high uncertainty due to the number of respondents and because not all animals are taken into account. Since the data are not available annually and incomplete, these data are only used as a check.

Combining data from the DIBEVO survey with the emission factors in Sutton et al (2000) results in a NH3 emission approximately 20% lower than the NH3 emission calculated with the emission factor of 0.2 kg per household. However, the DIBEVO survey only includes cats and dogs and the emission factors are derived from Sutton et al (2000).

In Sutton et al (2000), the uncertainty range provided is 50%. However, emission factors can be higher than this, according to the other studies.

The uncertainty in the activity data is small; rated as an A.

The emission factor depends on the share of different domestic animals, which varies in time. In addition, the emission factor is relatively uncertain, therefore the emission factor is qualified with an E.

The uncertainty of the total emissions is similar to the uncertainty in the emission factor. Therefore the uncertainty for the total emission is qualified as E.

Quality codes

Substance	Activity data	Emission factor	Emission
NH ₃	Α	E	E

Quality checks

No sector-specific quality checks were performed. For the general QA/QC, see chapter 2.

30.4 Spatial allocation

The emissions of ammonia from the manure of domestic animals are spatially allocated in the Netherlands based on inhabitants.

Emission source/process	Allocation-parameter	
Manure of domestic animals	Inhabitants	

Details available via:

http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation)

30.5 References

Booij H., 1995, Gezelschapsdieren, WESP-rapport C6, RIVM-rapport 772414003.

Bouwman et al, 1997, A global high-resolution emission inventory for ammonia, global biochemical cycles, Vol. 11, No. 4, pages 561-587

DiBeVo, https://dibevo.nl/

Feiten en cijfers gezelschapsdierensector 2015, August 2015, HAS Hogeschool

Joshua Fu et al, 2010, Quality Improvement for Ammonia Emission Inventory, Department of Civil and Environmental Engineering, University of Tennessee, Knoxville, TN 37996-2010

Neijenhuis F. and Niekerk T., 2015, 'Als de kat van huis is...', WUR report 316

NU.nl, http://www.nu.nl/algemeen/699151/nederlanders-houden-steeds-meer-huisdieren.html, 2006, viewed May 2015.

Sutton M.A. et al, 2000, Ammonia emissions from non-agricultural sources in the UK, Atmospheric Environment 34 (2000) 855-869

30.6 Version, dates and sources

Version: 1.0

Date: February 2018

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Pesticides, domestic and non-agricultural use

This section describes the emissions caused by the use of domestic pesticides; it excludes any agricultural use.

Process description	Emk_code	NFR_code	Sector
Domestic pesticides	0802400	2D3a	Consumers
Pesticides (non-	0812400	2D3i	Trade and
agricultural use)			Services

31.1 Description emission source

During domestic use of pesticides, emissions of NMVOC occur as a result of the propelling agent and/or other additives needed to distribute the pesticides in the appropriate dosage. Domestic pesticides are mainly used against unwanted indoor bugs (flies, mosquitos, ants etc.). The non-agricultural use of pesticides is mainly subscribed to the professional treatment of woodworm. The active pesticides are mixed with additives containing volatile organic compounds (NMVOC) which are emitted.

Pesticides used for agricultural purposes are reported by the Agriculture taskforce.

Contribution to the national emission

The contribution of NMVOC for non-agricultural and domestic pesticides use to the total national NMVOC emission is <1%.

31.2 Calculation

During the KWS2000 project, annual domestic and non-agricultural use of pesticides was estimated for the period 1990-2000. After 2001, the emissions were no longer updated.

For pesticides used for the pest control of woodworm, the estimate was based on information from the largest product resaler.

For domestic pesticides, the emission estimation was based on information provided by four companies responsible for the majority of the market share.

In 2017, TNO investigated possibilities to improve the NMVOC emissions in the WESP taskforce. However, no easy solution could be found to improve estimates of emissions for domestic and non-agricultural pesticide use due to the lack of activity data. Therefore, the emissions are still based on data from the KWS2000 project.

The NMVOC is divided into single substances according to emission profiles. For domestic pesticides, the NMVOC profile is based on the composition of propellant agents in aerosol cans. For the non-agricultural (woodworm) pesticides, the NMVOC emission profile is based on the permitted pesticide register ('Toelatingsregistratie Bestrijdingsmiddelen') (Klein, 1996).

31.3 Uncertainty

As the data are outdated, the uncertainty level for current emissions and the emission profiles is unclear. Up until 2000, the emission data and emission factors could be rated with an C. Currently, the emission data and emission factors should be rated with an E.

Substance	Activity data	Emission factors	Emission
NMVOS	В	С	С

31.4 Spatial allocation

The emissions of non-agricultural and domestic pesticides are regionalised in the Netherlands based on population density.

Source	Allocation-parameter	
Domestic pesticides	Population density	
Non-agricultural pesticides	Population density	

Details are available via:

http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation)

31.5 References

Dutch Emission Inventory.

http://emissieregistratie.nl/erpubliek/bumper.en.aspx

Klein A.E. 1996, Risico-evaluatie van hulpstoffen in nietlandbouwbestrijdingsmiddelen, TNO rapport R 96/319.

KWS2000, multiple years, Annual reports. InfoMil, Den Hague.

31.6 Version, dates and sources

Version 1.1

Date: February 2018

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32 Dry cleaning of clothing and textiles

This section describes the emissions caused by the dry cleaning of clothes and textiles. This section includes smaller consumer-focussed companies (<10 employees). Industrial cleaning of clothing and textiles is covered in section 33.

Process description	Emk_code	NFR_code	Sector
Chemical cleaning of	8922200	2D3f	Trade and
clothing and textile			Services

32.1 Description emission source

The dry cleaning of clothing and textiles leads to emissions of solvents. Hundreds of smaller companies (<10 employees) provide dry cleaning of clothing and textiles as service to consumers. The emissions calculated in this section only relate to the emissions of perchloroethylene (PER). Although in recent years the number of dry cleaning installations using PER has been declining, no alternative emission substances have been identified and/or calculated. Alternative solvents are, for example, CO2, petroleum-based hydrocarbons, or specialised detergents in water.

When the dry cleaning process is started, the clothing is put in a special washing machine containing PER and detergents. The PER is heated and vaporised; after cleaning the clean PER vapours are cooled and reused. A part of the PER forms a residue with the detergents and is wasted. A part of the PER vapour is emitted to the air. Another part of the PER goes into the 'contact' water, this contact water is filtered with active carbon filters and released to the sewage system.

Contribution to the national emission

The contribution of NMVOC for dry cleaning to the total national NMVOC emission to is <0.1%.

The contribution of PER for dry cleaning to the total national emission is about 20%.

32.2 Calculation

Emissions are calculated as follows:

Emission = Activity data x Emission factor

Activity data = Tonne of clothing and textile cleaned with PER Emission factor = kg emission per tonne dry cleaned

a) Activity data

Between 1994 and 1998, the amount of clothing was estimated based on inhabitants and the amount of clothing and textiles each inhabitant needed to have dry cleaned.

It is assumed that the activity data for 1990-1993 is equal to 1994. From 1999 onwards, the amount of clothing cleaned is based on the statistics from NETEX, a branch organisation for dry cleaning in the Netherlands. In their annual reports the distribution of the turn-over and the percentage of PER cleaning is mentioned, although not all years give

insights in the percentage of PER used as cleaning method. The use of PER as a cleaning method has declined over time. The turnover distribution is corrected for inflation using the data from Statistics Netherlands. For the years that no percentage of PER cleaning is mentioned in the NETEX annual report, the emissions have been calculated with an interpolated value of the percentage PER used.

b) Emission factor

Since 1980, there has been a constant push to reduce the emissions of PER from dry cleaning. First the machines had to be closed, later the vapours cooled and filtered. This has led to a reduction in the emission factor over time. In 1990, the emission factor started with 40 kilograms PER for each tonne of clothing and textiles cleaned. Between 2000 and 2007, the emission factor decreased from 40 kg PER to 20kg PER for each tonne of cleaned product. From 2007 onwards, the emission factor has been constant. Communication with NETEX in 2014 suggested that the emission factor could be as low as 10 kilograms. However, there was no reference to support this lower emission factor. Both the 40 kilogram PER and 20 kilogram PER for each tonne of clothing and textiles match the EMEP Guidebook emission factors.

32.3 Uncertainty

The activity data is calculated by using data from different sources, therefore the activity data is rated with an E.

The emission factors are based on reliable reports, but the market share of the technique used is unknown. Also, NETEX has doubts about the emission factor used. Therefore, emission factors are rated with a D. Overall the emissions are rated with an E.

Substance	Activity data	Emission factors	Emission
NMVOC/PER	E	D	E

32.4 Spatial allocation

The emissions of consumers are regionalised in the Netherlands based on population density.

Source	Allocation-parameter
Dry cleaning	Companies within SBI 93.01 (empl. <10)

Details are available via:

http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation)

32.5 References

EMEP/EEA air pollutant emission inventory guidebook 2016, 2.D.3.f Dry cleaning.

NETEX annual reports, multiple years. https://www.netex.nl/

32.6 Version, dates and sources

Version 1.1

Date: February 2018

Administrator	Organisation	E-mail address
B.I. Jansen	TNO	bart.jansen@tno.nl

33 Industrial cleaning of clothing and textiles

This section describes the emissions caused by cleaning clothes and textiles. This section includes industrial cleaning companies (>10 employees). Dry cleaning of clothing and textiles for consumers is covered in section 32.

Process description	Emk_code	NFR_code	Sector
Chemical cleaning of	8922100	2D3f	Trade and
clothing and textiles			Services

33.1 Description emission source

The industrial dry cleaning of clothing and textiles leads to emissions from solvents. The major companies considered for this emission source clean, for example, company clothing and cleaning cloths. The emissions calculated in this section contain the emissions of perchloroethylene (PER) and trichloroethene (TRI). Although in recent years the number of installations using PER and TRI are thought to be declining, no alternative emission substances have been identified and/or calculated. Alternative solvents used are, for example, CO₂, petroleum-based hydrocarbons or specialised detergents in water.

When the dry cleaning process starts, clothing is put in a special washing machine containing solvents (PER or TRI) and detergents. The PER is heated and vaporised, after the cleaning the clean PER vapours are cooled and reused. A part of the PER or TRI forms a residue with the detergents and is wasted. A part of the vapour is emitted to the air. Another part of the solvent goes into the 'contact' water, this contact water is filtered with active carbon filters and released to the sewage system.

Contribution to the national emission

The contribution of NMVOC for industrial cleaning to the total national NMVOC emission is <0.1%.

33.2 Calculation

Emissions are calculated as follows.

For the year 1991, the emissions of PER were calculated by the amount of PER sold to the market for cleaning purposes. This was (according to VHCP) an amount of 1690 tonnes. The amount of chemical waste containing PER was 600 tonnes, containing about 225 tonnes of PER. This resulted in a total emission of 1690 - 225 = 1465 tonnes of PER for industrial cleaning and dry cleaning for consumers together. In 1991, the PER emissions caused by dry cleaning for consumers was estimated at 965 tonnes. This results in an emission of 500 tonnes PER from the industrial cleaning of textiles and clothing.

All other estimates for this emission source were made based on expert judgement by H. vd Berg from TNO (industrial cleaning techniques). For 1995, the PER emission of industrial cleaning was estimated at 400 tonnes. The emissions of TRI were estimated at 11.9 tonnes in 1991 and

9.5 tonnes in 1995; no new estimates are available. Thus, the amount of emissions of both PER and TRI has been assumed to be constant from 1995 onwards.

Most of the PER and TRI are emitted to the atmosphere. The extra emissions to water are estimated to be about 0.015% of the emissions to air.

The emission estimates for industrial cleaning of clothing and textiles are outdated. The expectation is that the amount of PER and TRI used for cleaning has declined. It is also possible that the washing techniques have improved, resulting in less emissions.

33.3 Uncertainty

Since the data is outdated and based on expert judgement, the uncertainty is qualified with an E.

Substance	Activity data	Emission factors	Emission
All	E	E	Е

33.4 Spatial allocation

The emissions of industrial cleaning companies are regionalised in the Netherlands based on company location.

Source	Allocation-parameter
Dry cleaning	Companies within SBI 93.01 (empl. >10)

Details are available via:

http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation)

33.5 References

Berg H. van den, expert judgement by oral communication, TNO Cleaning techniques, Delft

CFK-project centre, 1991-1995, Tilburg

Dutch emission inventory,

http://emissieregistratie.nl/erpubliek/bumper.en.aspx

33.6 Version, dates and sources

Version 1.0

Date: February 2018

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Fumigation of transports

This section describes the emissions caused by fumigation of transports.

Process description	Emk_code	NFR_code	Sector
Fumigation of transports	E800000	2D3i	Trade and
			Services

34.1 Description emission source

Until 2008, containers and some bulk cargo ships were treated with methyl bromide for overseas transport. Treatment of overseas transport occurs in order to comply with the import legislations. For the Dutch market this is mainly done for grain, rice, cacao, nuts and animal feed. Fumigation occurs by putting pellets or tablets within the cargo container; during transport the humidity and oxygen react with the pellet to form toxic fumes containing methyl bromide. During transport and at the port, the methyl bromide is released to the air.

Contribution to the national emission

In the last reporting years, the only source of methyl bromide was the fumigation of transports.

34.2 Calculation

The emissions up to 2004 are based on numbers provided by the inspection of VROM (Ministry of housing, spatial planning and environment) called the 'Inspectie Milieuhygiëne' (IMH, stands for inspection environmental hygiene). The IMH checks incoming transports which have been chemically treated in order to supervise safe degassing of the transport in question. However, this is only the case for bulk transport. Container transport is not monitored, as it is not mandatory to report incoming containers which have been fumigated. Thus, only bulk transport was reported by IMH.

Between 2004 and 2008 no new data were provided by IMH. Since 2008, the use of methyl bromide has been prohibited. Although it is expected that other substance are used, no information can be found regarding amounts and substances used.

KPMG used to monitor the use of methyl bromide by questioning relevant companies with regards to the CFK action program. The data obtained by KPMG were higher than the data provided by IMH. This was due to illegal use; i.e. not reporting actual consumption to IMH as well as the lack of checks by IMH. However, the data provided by IMH are considered more consistent in time, so the KPMG data have not been used.

34.3 Uncertainty

As this emission source is outdated, it is hard to quantify the uncertainty.

However, due to the difference between the KPMG data and the IMH data, it is assumed that the uncertainty is high and rated with an E.

34.4 Spatial allocation

The emissions caused by the fumigation of transports were regionalised in the Netherlands based on the location where they take place.

Source	Allocation-parameter
Fumigation of transports	Location of activity

Details are available via:

http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation)

34.5 References

Dutch emission inventory,

http://emissieregistratie.nl/erpubliek/bumper.en.aspx

VROM, IMH, department pest control, written and oral communication.

34.6 Version, dates and sources

Version 1.1

Date: February 2018

Administrator	Organisation	E-mail address
B.I. Jansen	TNO	bart.jansen@tno.nl

35 Preserved wood

This section describes the emission from preserved wood.

Process description	Emkcode	NFR code	Sector
Solvent and other product use:		2D3i	Consumers
creosote pressure treated wood, new	0804000		
Solvent and other product use:		2D3i	Trade and
creosote pressure treated wood, new	0804001		services
Solvent and other product use:		2D3i	Agriculture
creosote pressure treated wood, new	0804002		
Solvent and other product use:		2D3i	Transport
creosote pressure treated wood, new	0804003		
Solvent and other product use:		2D3i	Consumers
creosote pressure treated wood, stock	0804100		
Solvent and other product use:		2D3i	Trade and
creosote pressure treated wood, stock	0804101		services
Solvent and other product use:		2D3i	Agriculture
creosote pressure treated wood, stock	0804102		
Solvent and other product use:		2D3i	Transport
creosote pressure treated wood, stock	0804103		
CCA pressure treated wood:		-	
constructions	0804200		
CCA pressure treated wood: garden		-	
funiture	0804201		
Leaching CCA pressure treated wood		-	
at waterline boarders, stock	0804300		

35.1 Description emission source

Emissions from preserved wood originate from three types of preserved wood:

- wood treated with Wolman salts (CCA pressure treated wood);
- wood treated with creosote (creosote pressure treated wood).
- carbolized wood (addressed separately in Chapter 11)

Wolman salts can cause emissions of arsenic, chromium and copper. Creosote preserved wood is a source of emissions of polycyclic aromatic hydrocarbons (PAH).

35.2 Calculation

Emissions are calculated for each substance by multiplying an activity rate (AR), in this case the quantity of preserved wood, by an emission factor (EF) expressed in kg of the specific substance per m² of preserved wood. This method of calculation is explained in the 'Handreiking Regionale aanpak diffuse bronnen'[1]. For creosote treated wood, this is calculated using the following equation:

Emission = AR_I x EF_I + Σ (AR)_S + EF_S

Where:

 AR_1 = Creosote treated wood (added last year) (m^2),

EFI = Emission factor creosote treated wood added last year

 (kg/m^2) AR $_S$ = Creosote treated wood (standing from previous years)

(m²)

EF_S = Emission factor creosote treated wood standing from

previous years (kg/m²)

A distinction is made between new application of creosote treated wood and standing quantities of creosote treated wood. "New application" refers to the amount of wood placed in the past year. "Standing" creosote treated wood refers to the amount of creosote treated wood placed in previous years and which is still a source of emissions.

For wood treated with Wolman salts, the emissions caused by standing wood from past years is calculated and totalled separately, as with the following formula:

Emission = Σ (AR J * EF J)

Where:

AR J = Amount of wood treated with Wolman salts placed in year J

(m³)

EF J = Emission factor for wood treated with Wolman salts placed in year J (kg/m^3) The emission calculated in this way is referred to as the total emission.

a) Activity data

The activity data is the amount of preserved wood. The activity data for creosote treated wood, wood treated with Wolman salts and carbolized wood are determined by different methods.

Creosote treated wood

For creosote treated wood, the assumption is that the emissions in the first year are higher than those in the years thereafter. This is why there are separate calculations for the new application (placed in the past year) and the standing (placed in previous years) wood.

The activity data are determined in the following manner:

The new application in 1985 and 1992 was determined by Hulskotte [2]. The quantities in 1990, 1995 and 2000 are interpolated. From 2001 on, no new creosote treated wood was placed, pursuant to the governmental regulation called PAK-besluit [3].

The standing wood in 1992 was calculated by multiplying the new application from 1980 (500,000 m2) by the lifetime (25 years) minus 10 years and then adding the new application of 1992 (250,000 m2) times 10 years [2]. Because we do not count the new application from the last year with the standing wood, we subtract the 1992 figure for new application from the standing wood calculated. This results in a figure of 9,750,000 m2 for standing wood 1992. This figure is used for the further

calculation of the standing wood in the other years. Figures for standing wood in other years is calculated in the same manner: Standing Wood reporting year = Standing Wood 1992 – a * New App 1980 + New App 1992 - reporting year

Where:

Standing Wood reporting year = Standing Wood in 1985, 1990,

1995, 2000, 2004 or 2005, (m2)

Standing Wood 1992 = Standing Wood in 1992, (9,750,000

m2)

a = Number of years between reporting

year and

1992

New App 1980 = New Application in 1980, (500,000

m2)

New App 1992—reporting year = Sum of the new application in the

years 1992 until reporting year,

(m2).

Wood treated with Wolman salts

Calculating the emissions by wood treated with Wolman salts requires knowing how much wood was placed in previous years. These quantities are based on [4]. Because the emission factor declines as the wood ages, counting the total amount of standing wood does not lead to a correct result. Instead, the emission per year of new application must be determined.

Wood treated with Wolman salts has only been used since 1979. In the years prior to that, creosote treated wood was most commonly used. For the years before 1979, the amount of Wolman salts treated wood placed is set at 0 m3. Its lifetime is set at 40 years [4]. From 2001 on, no new wood treated with Wolman salts was used in bank revetments, because no further WVO permits (permits under the Act on Water Pollution) were issued for the product after that time.

b) Emission factors

The emission factor is the emission per quantity of preserved wood in bank revetments. The emission factors for creosote treated and wood treated with Wolman salts are determined using different methods.

Creosote treated wood

For creosote treated wood, emission factors are formulated as being higher in the first year than in the years thereafter. This is why different emission factors are used for the new application of creosote treated wood and standing creosote treated wood. First, the emission factor for fluoranthene is determined. The emission factors for other substances are determined using a substance profile at leaching of the PAH. For new application of creosote treated wood, the emission factor for fluoranthene is calculated using the following assumptions:

Emission is highest for the first 31 days. For pine wood, the assumed emission factor is 4.0×10^{-6} kg fluoranthene/m²/day, and for fir, the assumed emission factor is 1.9×10^{-6} kg fluoranthene/m2/day [5]. For days 32-365 of the first year, the assumed emission factor is 0.9×10^{-6} kg

fluoranthene/m²/day for both woods [5]. Additionally, it is assumed that pine makes up approximately 75% of the wood used, with 25% being fir [6]. When combined, this information results in an emission factor of 4.1 * 10-4 kg fluoranthene/m²/year.

To calculate the emission factor for fluoranthene for standing creosote treated wood, the assumption is that this emission factor is the same as the emission factor for days 32-365 in the first year. This is an emission factor of 0.9 * 10-6 kg fluoranthene/m2/day for both woods [5]. This results in an emission factor of 3.3 * 10-4 kg fluoranthene/m²/year. The emission factors for phenanthrene, anthracene and pyrene are determined using data from a report by TNO [7]. This report presents the substance profile in the leaching fluids from two reports [5, 8]. Based on the substance profile, we estimate the ratios at leaching for phenanthrene, anthracene, fluoranthene and pyrene at 65%, 5%, 15% and 15% respectively.

The emission factor for naphthalene is taken from [9], in which the quantity of naphthalene is equal to the quantity of phenanthrene. This is why the same emission factor is used for these two substances.

Table 3: Emission factors for PAH-compounds from creosote treated wood, (10^{-3})

 kg/m^2) [5].

	New application	Standing
Phenanthrene	1.78	1.43
Anthracene	0.14	0.11
Fluoranthene	0.41	0.33
Pyrene	0.41	0.33
Naphthalene	1.78	1.43

Wood treated with Wolman salts

The emission factors for wood treated with Wolman salts depend on the type of preservative compound used to treat the wood and the leaching over the lifetime of the wood. More details about the calculation of the emission factors is available in [4]. Tables 4-6 show the emission factors for arsenic, chromium and copper, depending on the year of placement of the wood.

Table 4: Emission factors for arsenic from wood treated with Wolman salts based on [4]. (10-3 kg/m³)

Table 4: Emission		ctors for arsenic from wood treated with Wolman salts based on [4], (10-3 kg/m³)														
	Emiss	sion fa	ctor in	repor	ted yea	ar										
Year of new application	1985	1990	1995	2000	2005	2006	2007	2008	2009	2010	2011	2012	2013	2014	2015	2016
1979	8.19	7.61	7.15	6.83	6.44	6.44	6.4	6.3	6.2	6.2	6.1	6.0	6.0	5.9	5.9	5.9
1980	8.11	7.48	7.04	6.66	6.28	6.22	6.2	6.2	6.1	6.03	5.97	5.91	5.84	5.78	5.72	5.72
1981	7.95	7.28	6.86	6.49	6.13	6.07	6.0	6.0	5.9	5.88	5.82	5.76	5.70	5.64	5.58	5.52
1982	7.9	7.14	6.67	6.32	5.97	5.91	5.9	5.8	5.8	5.73	5.67	5.62	5.56	5.50	5.44	5.38
1983	7.83	6.99	6.53	6.14	5.8	5.75	5.7	5.6	5.6	5.58	5.52	5.46	5.41	5.35	5.30	5.24
1984	7.85	6.83	6.34	5.96	5.69	5.58	5.5	5.5	5.4	5.36	5.36	5.31	5.25	5.20	5.15	5.09
1985	8.37	6.71	6.19	5.82	5.51	5.46	5.4	5.3	5.3	5.20	5.15	5.15	5.10	5.04	4.99	4.94
1986		6.81	6.24	5.88	5.56	5.51	5.5	5.4	5.3	5.25	5.20	5.15	5.15	5.10	5.04	4.99
1987		7.02	6.34	5.93	5.62	5.56	5.5	5.5	5.4	5.30	5.25	5.20	5.15	5.15	5.10	5.04
1988		7.23	6.45	6.03	5.67	5.62	5.6	5.5	5.5	5.36	5.30	5.25	5.20	5.15	5.15	5.10
1989		7.54	6.55	6.08	5.72	5.67	5.6	5.6	5.5	5.46	5.36	5.30	5.25	5.20	5.15	5.15
1990		4.51	3.61	3.33	3.14	3.08	3.1	3.0	3.0	2.97	2.94	2.88	2.86	2.83	2.80	2.77
1991			3.67	3.36	3.16	3.14	3.1	3.1	3.0	3.00	2.97	2.94	2.88	2.86	2.83	2.80
1992			3.78	3.42	3.19	3.16	3.1	3.1	3.1	3.02	3.00	2.97	2.94	2.88	2.86	2.83
1993			3.89	3.47	3.25	3.19	3.2	3.1	3.1	3.05	3.02	3.00	2.97	2.94	2.88	2.86
1994			3.05	2.65	2.46	2.44	2.4	2.4	2.4	2.31	2.29	2.27	2.25	2.23	2.21	2.16
1995			3.1	2.48	2.29	2.25	2.2	2.2	2.2	2.16	2.12	2.10	2.08	2.06	2.04	2.02
1996				2.29	2.1	2.08	2.0	2.0	2.0	1.98	1.96	1.93	1.91	1.89	1.87	1.86
1997				2.13	1.92	1.89	1.9	1.8	1.8	1.80	1.78	1.76	1.73	1.72	1.70	1.69
1998				1.95	1.74	1.71	1.7	1.7	1.6	1.62	1.60	1.58	1.57	1.54	1.53	1.51
1999				1.69	1.47	1.45	1.4	1.4	1.4	1.37	1.35	1.33	1.32	1.31	1.28	1.27
2000				1.5	1.2	1.18	1.2	1.1	1.1	1.11	1.09	1.08	1.06	1.05	1.05	1.03
2001-2016					0	0	0	0	0	0	0	0	0	0	0	0

Table 5: Emission factors for chromium from wood treated with Wolman salts based on [4], (10^{-3} kg/m^3)

Year of new	1985	1990	1995	2000	2005	2006	2007	2008-2017
application								
1979	0.200	0	0	0	0	0		
1980	0.202	0	0	0	0	0		
1981	0.204	0	0	0	0	0		
1982	0.206	0	0	0	0	0		
1983	0.208	0.208	0	0	0	0		
1984	0.420	0.210	0	0	0	0		
1985	1.855	0.212	0	0	0	0		
1986		0.212	0	0	0	0		
1987		0.212	0	0	0	0		
1988		0.212	0.212	0	0	0		
1989		0.424	0.212	0	0	0		
1990		2.065	0.236	0	0	0		
1991			0.236	0	0	0		
1992			0.236	0.000	0	0		
1993			0.236	0.236	0	0		
1994			0.432	0.216	0	0		
1995			1.523	0.218	0	0		
1996				0.219	0	0		
1997				0.221	0	0		
1998				0.222	0.222	0		
1999				0.431	0.215	0.215		
2000				1.461	0.209	0.209	0.209	0
2001-2016					0	0	0	0

							2		
Tabla 4.	Emission	faatara fa	raannar	from wood	trooted	with Malman	00/10/10-1	1/02/02.31	based on [4].
TADIE O'	rmssion	$1acioi \le 10$	i coodei	11 ()[[] VV()()(]	niearea v	viin vvoiman		K(1/111°)	08500 00 141

Table 6: Emission		sion fa						rr surts	(,,,	<u> </u>	Buscu		•			
Year of new application	1985	1990	1995	2000	2005	2006	2007	2008	2009	2010	2011	2012	2013	2014	2015	2016
1979	2.0	1.2	0.8	0.6	0.4	0.4	0.4	0.2	0.2	0.2	0.2	0.2	0.2	0.2	0.2	0.2
1980	2.2	1.4	8.0	0.6	0.4	0.4	0.4	0.4	0.2	0.2	0.2	0.2	0.2	0.2	0.2	0.2
1981	2.4	1.4	1.0	0.6	0.4	0.4	0.4	0.4	0.4	0.2	0.2	0.2	0.2	0.2	0.2	0.2
1982	2.6	1.6	1.0	0.6	0.4	0.4	0.4	0.4	0.4	0.4	0.2	0.2	0.2	0.2	0.2	0.2
1983	2.8	1.8	1.2	8.0	0.6	0.4	0.4	0.4	0.4	0.4	0.4	0.2	0.2	0.2	0.2	0.2
1984	3.4	2.0	1.2	0.8	0.6	0.6	0.4	0.4	0.4	0.4	0.4	0.4	0.2	0.2	0.2	0.2
1985	22.0	2.2	1.4	0.8	0.6	0.6	0.6	0.4	0.4	0.4	0.4	0.4	0.4	0.2	0.2	0.2
1986		2.4	1.4	1.0	0.6	0.6	0.6	0.6	0.4	0.4	0.4	0.4	0.4	0.4	0.2	0.2
1987		2.6	1.6	1.0	0.6	0.6	0.6	0.6	0.6	0.4	0.4	0.4	0.4	0.4	0.4	0.2
1988		2.8	1.8	1.2	0.8	0.6	0.6	0.6	0.6	0.6	0.4	0.4	0.4	0.4	0.4	0.4
1989		3.4	2.0	1.2	0.8	0.8	0.6	0.6	0.6	0.6	0.6	0.4	0.4	0.4	0.4	0.4
1990		17.6	1.8	1.1	0.6	0.6	0.6	0.5	0.5	0.5	0.5	0.5	0.3	0.3	0.3	0.3
1991			1.9	1.1	8.0	0.6	0.6	0.6	0.5	0.5	0.5	0.5	0.5	0.3	0.3	0.3
1992			2.1	1.3	8.0	8.0	0.6	0.6	0.6	0.5	0.5	0.5	0.5	0.5	0.3	0.3
1993			2.2	1.4	1.0	8.0	0.8	0.6	0.6	0.6	0.5	0.5	0.5	0.5	0.5	0.3
1994			2.6	1.5	0.9	0.9	0.8	8.0	0.6	0.6	0.6	0.5	0.5	0.5	0.5	0.5
1995			5.2	1.8	1.1	1.0	1.0	8.0	0.8	0.6	0.6	0.6	0.5	0.5	0.5	0.5
1996				2.0	1.2	1.2	1.0	1.0	8.0	0.8	0.7	0.7	0.6	0.5	0.5	0.5
1997				2.2	1.3	1.2	1.2	1.0	1.0	0.8	0.8	0.7	0.7	0.7	0.5	0.5
1998				2.5	1.5	1.4	1.2	1.2	1.0	1.0	0.8	0.8	0.7	0.7	0.7	0.5
1999				3.2	1.7	1.5	1.4	1.2	1.2	1.0	1.0	8.0	8.0	0.7	0.7	0.7
2000				6.4	1.9	1.7	1.5	1.4	1.2	1.2	1.0	1.0	0.8	8.0	0.7	0.7
2001-2016					0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0

Until 1989, treatment compound CCA type B was used. In 1990, this was replaced by preservative type C [4] (see appendix 1). Preservative CCA type C contains less arsenic than CCA type B. The replacement of Type B with Type C resulted in an emission reduction of arsenic.

From 2001 on, no new creosote treated wood was placed, pursuant to the governmental regulation called PAK-besluit [3]. Likewise, from 2001 on, no new wood treated with Wolman salts was used in bank revetments because no further WVO permits (permits under the Act on Water Pollution) were issued for the product after that time. Application of preserved wood in, along or above water is obliged to obtain a WVO permit. One consideration in any permitting procedure is that creosote treated wood and wood treated with Wolman salts are a source of environmental problems, even though alternatives are available. Consequently, these wood preservatives are no longer allowed in bank revetments.

From 2001 on, only emission occurs from preserved wood that was placed prior to that year.

Release into environmental compartments

The assumption for wood treated with Wolman salts is that all emissions go directly to surface water [4]. For creosote treated wood, the assumption is that half of the wood comes into direct contact with water, and consequently that half of the emissions goes to the soil and half to the surface water [5].

35.3 Uncertainty

The activity data for creosote treated wood, wood treated with Wolman salts are both based on extrapolation of estimates. This is assigned a classification of D. The emission factors are determined using measurements supplemented with assumptions and estimates. This is assigned a classification of C (for creosote treated, wood treated with Wolman salts).

35.4 Spatial allocation

The emissions of preserved wood are allocated in the Netherlands based on length of bank revetments. Details are available at http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation)

35.5 References

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- Rijsdijk, J.F. en J.D.P. de Ruiter. Uitloging van gecreosoteerd hout. VHN, rapport 1,2 en 3.
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35.6 Version, date and sources

Version: 1.0 Contact:

Emission expert	Organisation	E-mail adress
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36 Refrigerator foam

This section describes the emissions caused by CFC in refrigerator foam.

Process description	Emk_code	NFR_code	Sector
Diffuse emission insulation foam refrigerator	0890400	2D3i	Waste disposal
Diffuse emission insulation foam refrigerator	0890401	2D3a	Consumers

36.1 Description emission source

When refrigerator foam is made with CFCs as blowing agent, the CFCs are released during the use and waste of the refrigerator. Hence since the Montreal protocol prohibited the use of CFCs, the production of foam with CFC-11 has ceased. Since 1995, no new refrigerators with CFC-11 insulation foam have been produced for the Dutch market. The emissions are distributed evenly between the sectors consumers and waste disposal. From 2013 on, the emissions are supposed to have been reduced to zero.

Contribution to the national emission

The contribution is no longer relevant as no emissions are possible.

36.2 Calculation

Emissions are calculated as follows: Emission = Activity data x Emission factor

Activity data = Number of refrigerators Emission factor = kg emission per refrigerator

a) Activity data

The number of refrigerators is based on the year 1992, in which a total of 6.35 million refrigerators was estimated. From 1994, the number of refrigerators containing CFCs has declined by 350,000 annually. This has resulted in zero refrigerators in 2013.

b) Emission factor

The emission factor is based on the year 1992. The emission calculated for this year was divided by the total number of refrigerators. This results in an emission to air of 17 grams trichlorofluoromethane per refrigerator. This emission factor was distributed between both emission sources, giving an emission factor of 8.5 gram trichlorofluoromethane per refrigerator per emission source.

36.3 Uncertainty

The activity data was based on a single report from 1995, thus the uncertainty is high, especially for later years, resulting in an E

The emission factor is also based on the same report. Together with the amount of CFC diffusing from foam declining over time; these factors make the uncertainty high and are rated with an E.

Substance	Activity data	Emission factors	Emission
All	Е	E	Е

36.4 Spatial allocation

The diffuse emissions of refrigerators are regionalised in the Netherlands based on population density.

Source	Allocation-parameter
Refrigerator diffuse emissions from	Population density
insulation foam	·

Details are available via:

http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation)

36.5 Reference

Brouwer J.G.H. et al, 1995, Verwerking afgedankte koelapparatuur, WESP-report H-2, RIVM report 772414004, Bilthoven.

36.6 Version, dates and sources

Version 1.1

Date: February 2018

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37 Tanker truck cleaning

This section describes the emissions of substances from tanker truck cleaning.

Process description	EMK code	Target group	NFR category
Tanker truck	0811200	Trades and	2D3i
cleaning		services	

37.1 Description emission source

The process is defined as the cleaning of tanker trucks by specialised cleaning companies (20 in number) united in the ATCN. There is a large variation in the loads carried by tanker trucks, e.g. orange juice, chalk powder, formaldehyde, glycol, phosphoric acid, natron leach, kerosene, wine, etc. In many cases the tanks still contain vapour or a rest load of volatile substances which are released during cleaning. This was the case for 41,000 tanker trucks in 1999.

37.2 Calculation

Emissions from cleaning tanker trucks were measured by TNO in 1999 and reported in 2000 (TNO, 2000). All post-1999 annual emissions have been set equal to these measured emissions. Annual emission estimations for the years prior to 1999 are based on estimations from the tanker truck branch and the KWS project (Infomil, 2002).

In a meeting held in 2000 with the tanker truck branch, the steps to be taken were discussed. The meeting's outcome led to some interventions to reduce emissions; these are reported in the NIR of 2001. However, it is unknown to what extent these interventions have been implemented. A letter was sent to the ATCN in 2005, but no reaction was received. Currently, emissions are assumed to be equal to those of the year 2000.

37.3 Uncertainty

Substance	Activity data	Emission factors	Emission
NMVOC			D

37.4 Spatial allocation

Spatial allocation of emissions is based on population density.

37.5 References

Infomil, 2002. KWS 2000 eindrapportage, Infomil, Den Haag 52TNO. 2000. Vervolgonderzoek naar emissies van VOS bij tankautoreiniging in Nederland, sept 2000, TNO-MEP r2000/280,

Apeldoorn

37.6 Version, date and sources

Version: 1.0

Date:

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38 Accidental fires

This section describes the emissions caused by accidental fires of cars and houses.

Process description	Emk_code	NFR_code	Sector
Accidental fires, houses	0801200	5E	Consumers
Accidental fires, cars	0801300	5E	Consumers

38.1 Description emission source

Due to accidents and/or purposeful actions, cars and houses are lost in a fire. The smoke resulting from the fire is the source of emissions. When a house or car burns, the amount of material lost in the fire is dependent on the response time of (professional) fire fighters.

Contribution to the national emission

The contribution of particulate matter for accidental fires to the total national PM_{10} emission is <0.1%.

38.2 Calculation

Emissions are calculated as follows: Emission = Activity data x Emission factor

Activity data = Number of accidental fires (house or car) Emission factor = kg emission per accidental fire

a) Activity data

The number of houses and cars exposed to fire was reported annually by CBS Statline until 2013. These data were used for the time series 1990-2013.

For the house fires after 2014, an assumption was made that 0.2% of houses are exposed to a fire.

For car fires, a news article was found giving the number of car fires for 2015 and 2016. This article refers to 'alarmeringen.nl' and appears legitimate. The year 2014 was interpolated from 2013 and 2015.

b) Emission factor

For car fires, emission factors were derived from the EMEP/EEA guidebook (EMEP/EEA, 2016) (chapter 5.E. table 3.2).

The emission factor for house fires in the EMEP/EEA guidebook (chapter 5.E. table 3.3 till 3.5) seems inappropriate for the Dutch situation. The emission factor in the guidebook is based on a Norwegian study, however Norwegian houses contain more wood and Norway is more rural.

To get an estimate for the Dutch situation, a study on the Dutch housing stock by TNO was used to derive an amount of combustible materials, leaving out the 95% non-combustible materials like concrete and bricks. Without the interior of the house, this results in about 10.3 tonnes. Based on expert judgement, the interior is estimated at 3 tonnes, making a total of 13.3 tonne. According to statistics from the Dutch fire

brigade (brandweerstatistiek 2013), about 90% of the fires are regarded as small and 10% as large. To correct for the amount of material destroyed in the fire, for small fires it is assumed that 25% of the material is combusted, and for large fires, 100% of the material is combusted. Therefore, on average, 32.5% of the house is lost in a fire, resulting in about 4.3 tonnes of materials burned. An emission factor of 2.5 g PM_{10} /kg was applied (based on open wood burning), resulting in 10.8 kg PM_{10} /fire.

The PM_{10} emission factor was converted to a $PM_{2.5}$ and EC emission factor based on expert judgement.

The PCDD/F emission factor is based on the relation between PCDD/F and PM_{10} emission factors presented in the EMEP/EEA guidebook. The emission factors of heavy metals presented for house fires in the guidebook are considered irrelevant for the Dutch situation, since the amounts released are very low (<<0.1%).

Cubatanaa	Houses		Cars	
Substance	EF	Unit	EF	Unit
PM ₁₀	10.8	kg/ fire	2.3	kg/ fire
PM _{2.5}	10.3	kg/ fire	2.3	kg/ fire
EC	2.1	kg/ fire	1.15	kg/ fire
PCDD/F (i-TEQ)	108	μg/ fire	48	μg/ fire

38.3 Uncertainty and quality checks

The uncertainty in the activity data for the years before 2014 is relatively low as the data are reported by a reliable source. From 2014 onwards, the activity data is not reported annually or had to be estimated. Therefore the uncertainty is rated with a C.

The emission factor for car fires is reported within the guidebook, with a relatively high bandwidth with only 1 source mentioned. Therefore the emission factor is rated with a B.

For house fires, the emission factors are based on expert judgment in combination with reliable sources. Therefore the emission factor is rated with a D.

Source	Activity data	Emission factors	Emission
House fires	С	D	D
Car fire	С	В	С

Quality checks

The number of house fires reported by the CBS is not in-line with the number of house fires reported in a report from the Dutch association of insurance companies. The insurance companies report about 7 times more house fires. However, it is believed that the insurance companies report every fire-related incident, for example an incident with burn stains in a kitchen.

38.4 Spatial allocation

on population density.

Source	Allocation-parameter
Accidental fires, houses and cars	Population density

Details are available via:

http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation)

38.5 References

Brandweerstatistiek 2013

CBS statline, https://opendata.cbs.nl/statline/#/CBS/en/

EMEP/EEA air pollutants emission inventory guidebook, 2016,

 $\underline{\text{https://www.eea.europa.eu/publications/emep-eea-guidebook-}}$

2016

Verbond van Verzekeraars, https://www.verzekeraars.nl/publicaties/actueel/risicomonitor-woningbranden-2017

38.6 Version, dates and sources

Version 1.1

Date: February 2018

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39 Degreasing new vehicles

This section describes the emissions caused by degreasing new vehicles. These emissions occurred until 1999.

Process description	Emk_code	Sector
Degreasing of new	8920701	Trade and Services
vehicles		

39.1 Description emission source

Until 1998, cars were sometimes protected with a paraffin coating (wax) to protect them during transport. At the dealer, the protective coating was removed with a solvent, resulting in emissions of NMVOC. Currently cars are protected by a plastic film or a water solvable coating. However, some car producers apply no protection at all, or only for a certain type of transport.

Contribution to the national emission

This emission source is no longer relevant.

39.2 Calculation

a) Activity data

Prior to 1999, emissions were calculated based on the number of imported cars (of certain brands) that were paraffin coated and dewaxed at the dealer. Most car brands had central depots with adequate air cleaning installations, so no significant emissions occurred. Between 1990 and 1993 fewer cars were dewaxed each year. As mentioned above, other options (centralized cleaning, other coatings) reduced the need for dewaxing a new car. Between 1994 and 1998, only one car brand used decentralized dewaxing.

b) Emission factors

An emission factor to air of 0.64 kg NMVOC per cleaned car was applied for each year. The emission factor to the sewer system was about 4.5 g NMVOC and the direct emissions to water were estimated to be 0.3 g NMVOC per treated car. These emission factors were estimated by TNO based on the reference year 1992. An emission profile for the NMVOC was also determined by TNO; the NMVOC emissions were supposed to be non-aromatic hydrocarbons.

39.3 Uncertainty

As this method is no longer used, it is hard to quantify the uncertainty. The TNO report on the emission profile is not (publicly) available, therefore it is difficult to qualify the value of the emission factors. If the uncertainty had to be rated, it would be qualified with an E.

39.4 Spatial allocation

The emissions of consumers are regionalised in the Netherlands based on population density.

Source	Allocation-parameter
Degreasing of new vehicles	Car maintenance companies

Details are available via:

http://www.emissieregistratie.nl/erpubliek/misc/documenten.aspx?ROO T=Algemeen (General)\Ruimtelijke toedeling (Spatial allocation)

39.5 References

CREM, November 1998, 'Monitoring VOS-emissie van 9 sectoren 1997', Amsterdam (not publicly available).

InfoMil, KWS2000 Annual reports, multiple years, Den Hague.

39.6 Version, dates and sources

Version 1.0

Date: February 2018

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40 Quality indication

The quality and reliability of the emission data is expressed in a coding system using A, B, C, D and E scores. This corresponds to the method used in EPA emission inventories in the light of EMEP/CORINAIR. The quality scores are defined as follows:

- A. The data are gathered from very accurate (high precision) measurements.
- B The data are gathered from accurate measurements.
- C The data are gathered from a published source such as government statistics or industrial trade figures.
- D The data are derived from extrapolation of other measured activities.
- E The data are derived from extrapolation of foreign data.
- N Not applicable or no data available.

The reliability of the emission factors can vary substantially over time and between substances. Therefore, no confidence interval can be linked to the quality indications used. However, it can be assumed that, for a specific substance, the relative confidence declines along the data quality classifications A to E.